

# Simulating carbon and nitrogen cycling in microplastic contaminated agroecosystems using gradient boost regression model

Shahid Iqbal<sup>1,2\*</sup>, Fiona Ruth Worthy<sup>3,4</sup>, Heng Gui<sup>1\*</sup> and YunJu Li<sup>1,2</sup>

<sup>1</sup> Department of Economic Plants and Biotechnology, Yunnan Key Laboratory for Wild Plant Resources, Kunming Institute of Botany, Chinese Academy of Sciences, Kunming, Yunnan 650201, China

<sup>2</sup> Honghe Centre for Mountain Futures, Kunming Institute of Botany, Chinese Academy of Sciences, Honghe, Yunnan 654400, China

<sup>3</sup> Key Laboratory of Phytochemistry and Natural Medicines, Kunming Institute of Botany, Chinese Academy of Sciences, Kunming, Yunnan 650201, China

<sup>4</sup> Yunnan Key Laboratory for Fungal Diversity and Green Development, Kunming Institute of Botany, Chinese Academy of Sciences, Kunming, Yunnan 650201, China

\* Correspondence: [shahiduaif85@gmail.com](mailto:shahiduaif85@gmail.com) (Iqbal S); [guiheng@mail.kib.ac.cn](mailto:guiheng@mail.kib.ac.cn) (Gui H)

## Abstract

Microplastics (MPs) pollution poses significant threats to carbon (C) and nitrogen (N) cycling, affecting plant biomass. These threats are often difficult to measure due to the multifaceted effects of MPs. Hence, machine learning models offer a promising approach to effectively estimate nutrient dynamics. Here, we employed a Gradient Boosting Regression (GBR) model to estimate soil C and N contents, greenhouse-gas emissions, and plant biomass in MPs-contaminated soils. We also evaluated the mediation of soil type and incubation duration. To train and test the GBR model, we used data compiled from 55 peer-reviewed publications. The model results showed strong agreement between observed and predicted values for dissolved organic carbon (DOC). Training and testing  $R^2$  values differed by 3% for  $\text{NH}_4^+$  and 2% for  $\text{NO}_3^-$ . For  $\text{CO}_2$  and  $\text{N}_2\text{O}$  emissions, low mean squared error values indicated strong model performance in predicting emissions under MPs pollution. The model demonstrated strong predictive performance for biomass, achieving high  $R^2$  values for both the training and testing datasets. Across properties, MPs size caused the greatest changes in DOC and  $\text{CO}_2$  emissions. However, the MP shape had the greatest impact on the SOM content (60%). The greatest changes in soil  $\text{NH}_4^+$  (36%) and  $\text{NO}_3^-$  (51%) content and plant biomass (77%) were caused by MP type. For  $\text{N}_2\text{O}$  emissions, the size, type, dose, and incubation period of MPs had substantial effects. Our results conclude that the GBR model is a powerful tool to estimate the effects of MPs pollution on nutrient cycling and plant biomass.

**Citation:** Iqbal S, Worthy FR, Gui H, Li Y. 2026. Simulating carbon and nitrogen cycling in microplastic contaminated agroecosystems using gradient boost regression model. *Circular Agricultural Systems* 6: e005 <https://doi.org/10.48130/cas-0026-0007>

## Introduction

Microplastic pollution (MP) is a global concern for nutrient cycling in agroecosystems<sup>[1–4]</sup>. Thus, nutrient cycling has garnered significant research attention as it provides insights into ecological stability and ecosystem functioning<sup>[2,5]</sup>. In this context, studying the effects of MPs on carbon (C) and nitrogen (N) cycling is particularly crucial, given their negative impacts on crop production and environmental health. By definition, MPs are plastic particles < 5 mm that may originate as primary micro-sized particles or result from the degradation of large plastic materials<sup>[6]</sup>. Among different environments, soils are the largest sink of MPs. Some agricultural soils have become permanent hotspots of MPs, receiving annually  $50 \times 10^4$  t MPs<sup>[7]</sup>. According to an estimate, plastic production will rise and subsequently may accumulate 50 million metric tons of plastic waste in our environment. After degradation, this plastic waste will contribute to MPs pollution. Given the considerable pressure of MP pollution in agroecosystems, it is important to mechanistically understand C and N cycling<sup>[2]</sup>.

MPs affect C and N cycling through a range of biochemical pathways. These pathways include alteration in microbial enzymes as well as changes in microbial diversity, abundance, and activities<sup>[8–11]</sup>. The presence of MPs can also indirectly alter C and N cycling by inducing soil physical and chemical changes, including bulk density, aggregation, pH, and organic matter<sup>[12–14]</sup>. However, these effects of MPs are highly variable and depend on a wide range of factors<sup>[2,15,16]</sup>. To date, a myriad of studies have been limited to

single factors and greenhouse studies. As a result, our mechanistic understanding of the underlying factors driving carbon (C) and nitrogen (N) cycling remains limited<sup>[15]</sup>. The involvement of a wide range of interacting factors makes the development of such a mechanistic understanding a significant challenge.

Given the diverse properties of MPs and their long residence time, research results, particularly regarding C and N cycling, are often inconsistent. In this context, machine learning (ML) models may offer a promising approach to addressing the complexity of MP contamination and accurately predicting carbon (C) and nitrogen (N) cycling<sup>[1]</sup>. Most importantly, ML models can effectively predict complex relationships between independent and dependent variables<sup>[17]</sup>. Recently, Withana et al.<sup>[1]</sup> used ML models to demonstrate that soil physicochemical properties are distinctly affected by MPs properties. However, there is still a paucity of MPs studies employing ML models<sup>[1]</sup>. Thus, we aimed to extract data from previous research and predict C and N cycling using a gradient boost regression (GBR) model. We hypothesized that the GBR model is an effective tool for predicting C and N cycling under varying properties of MPs. Our specific objectives were:

- (1) To estimate the impact of key properties of MPs on different forms of soil C and N contents,
- (2) To evaluate  $\text{CO}_2$  and  $\text{N}_2\text{O}$  emissions from MPs-polluted soils, and
- (3) To investigate how MPs pollution influences crop biomass.

The results of this study make an important contribution to the understanding of the interactions between the properties of MPs

and C and N cycling in agroecosystems. This knowledge will help inform decisions aimed at mitigating the impacts of MP pollution in these systems.

## Materials and methods

### Data retrieval

A literature search on the impact of MPs on soil C and N cycling was conducted using Google Scholar, PubMed, and Web of Science. The literature search identified a total of 450 peer-reviewed papers, published between 2017 and 2025. From the identified papers, 55 studies were selected that specifically assessed the effects of MPs on soil C and N content, CO<sub>2</sub> and N<sub>2</sub>O emissions, and plant biomass. Numerical data from the tables and figures in the selected papers were extracted. To extract data from the figures in the selected papers, we employed Web Plot Digitizer Software version 2.26. Additionally, information on experimental conditions such as soil type, experiment duration, MPs size, MPs shape, MPs type, and MPs dose was also extracted. The data were further screened to avoid duplication.

### Study design

We designed a predictive study using data extracted from previous publications to investigate the effects of MPs pollution on C and N cycling. The availability of substantial data on various properties of MPs and related factors facilitated the development of a machine learning model. Detailed information on the input variables is provided in Table 1. Soil chemical properties and greenhouse gas emissions such as dissolved organic carbon (DOC), soil organic carbon (SOC), soil organic matter (SOM), ammonium (NH<sub>4</sub><sup>+</sup>), nitrate (NO<sub>3</sub><sup>-</sup>), carbon dioxide (CO<sub>2</sub>) and nitrous oxide (N<sub>2</sub>O), as well as crop biomass were rigorously analyzed, and used as key candidates for training and testing the GBR model. All parameters were compiled separately and used to develop individual prediction models for the same type of input variables<sup>[1]</sup>. As all parameters exhibited unbalanced distributions, we performed a log transformation to reduce the data gap<sup>[1,18]</sup>. This step improved the normality of the data, as well as model convergence and performance.

### Model training and testing

We used the Gradient Boosting Regression (GBR) tree-based algorithm to develop the model using a boosting strategy. The GBR model was trained and tested using the retrieved data on C and N cycling<sup>[1]</sup>. The dataset was randomly partitioned into training and testing sets in an 80:20 ratio<sup>[1,19]</sup>. The training set (80%) was used for hyperparameter tuning and model training, while the testing set (20%) was used to evaluate the prediction performance of the trained model.

**Table 1.** Details of variables used for model development.

MPs properties	MPs type	PE, HDPE, PS, PLA, LDPE, PU, PVC, PA, PBS, PHB, PTFE, PP, PET, PAN, PBAT, PHE
	MPs size (µm)	1, 3, 5, 6, 13, 18, 25, 30, 35, 55, 57, 60, 66, 67, 100, 103, 125, 150, 180, 188, 190, 250, 500, 625, 630, 900, 1,000, 2,000, 4,280
	MPs shape	Particle, fragment, pellet, round, bead, fiber, powder, granule, vessel
	MPs dose (%)	0.001, 0.002, 0.01, 0.02, 0.05, 0.1, 0.2, 0.3, 0.25, 0.4, 0.5, 1, 1.5, 2, 2.5, 3, 5, 7, 10, 10.5, 13, 28
Soil related	Soil type	Clay loam, clay, sandy loam, loamy, sandy, sandy clay loam, silt loam, silty, loam, clay loam, silty clay loam
Experiment duration	Incubation (days)	7, 10, 12, 14, 15, 19, 21, 20, 22, 25, 28, 30, 31, 35, 40, 42, 45, 46, 50, 54, 55, 60, 70, 80, 90, 100, 105, 110, 120, 144, 164, 365

PE-polyethylene. HDPE-High-density polyethylene. PS-polystyrene. PLA-poly-lactic acid. LDPE-low-density polyethylene. PU-polyurethane. PVC-polyvinyl chloride. PA-polyamide. PBS-polybutylene succinate. PES-Polyester. PET-polyethylene terephthalate. PP-polypropylene. PHB-Polyhydroxybutyrate. PTFE-Polytetrafluoroethylene. PAN- Polyacrylonitrile. PBAT- Polybutylene adipate terephthalate.

For hyperparameter tuning, the maximum tree depth, number of trees, learning rate, and subsampling rate were adjusted to optimize the model for the compiled dataset. A five-fold cross-validation method was employed to prevent overfitting due to inappropriate hyperparameters. The optimal hyperparameters were determined based on the average performance across the validation folds. The coefficient of determination (R<sup>2</sup>) and root mean square error (RMSE) were used to evaluate the prediction accuracy of the trained model<sup>[20]</sup>.

### Statistics and data visualization

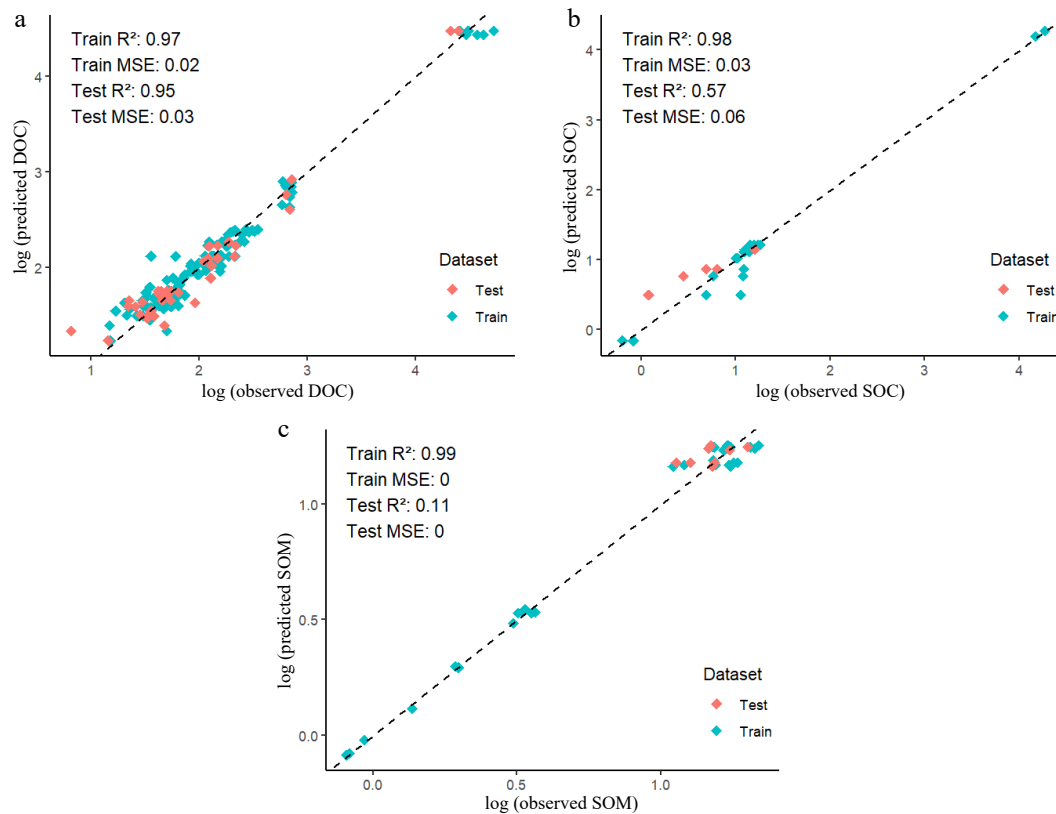
The correlation between observed and predicted data was assessed. All data visualizations were performed using RStudio (v4.2.2) and R (v4.5.1). The 'ggplot2' package was used to generate scatter plots with a 1:1 reference line to determine the deviation of observed and predicted values<sup>[21]</sup>. The 'annotate' function was employed to display R<sup>2</sup> and MSE values on each plot. Additionally, the 'radarchart' package was used to visualize the contribution of MP characteristics to changes in soil nutrient dynamics and plant biomass (R Development Core Team, 2008).

## Results and discussion

### Model evaluation

The GBR model demonstrated high accuracy in predicting various forms of soil carbon following MP pollution (Fig. 1). There was good agreement between the observed and predicted values of DOC (Fig. 1a). A small gap between training and testing performance for DOC suggests robust and reliable predictions. These results are consistent with the findings of Withana et al.<sup>[1]</sup>, who observed a small difference in R<sup>2</sup> between the trained and tested datasets for DOC. In contrast, a larger gap in R<sup>2</sup> values between training and testing for SOC and SOM indicates potential overfitting (Fig. 1b, c). These results indicate that the model was unable to fully capture dataset-specific patterns in SOC and SOM. This limitation likely reflects heterogeneity introduced by the use of diverse soil carbon measurement methods within the dataset. As a result, the model's ability to generalize to unseen SOC and SOM data collected through heterogeneous measurement approaches could be limited. Future studies could mitigate these limitations by increasing sample size and explicitly accounting for variability associated with soil carbon measurement techniques. While a relatively large difference in R<sup>2</sup> values was observed, the low mean squared error (MSE) values and the close alignment of predicted and observed values along the 1:1 line suggest that overall prediction accuracy remained reasonable.

The prediction of NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> content in MP-polluted soils was highly accurate (Fig. 2a, b). The difference between training and



**Fig. 1** Observed vs predicted content of (a) dissolved organic carbon (DOC), (b) soil organic carbon (SOC), and (c) soil organic matter (SOM) in microplastic polluted soils.

testing  $R^2$  values was only 3% for  $\text{NH}_4^+$  and 2% for  $\text{NO}_3^-$ . Moreover, the low MSE values and the close alignment of predicted and actual values along the 1:1 line for both training and testing datasets indicate strong model performance. Similarly, a previous study found that the GBR model can accurately predict plant available nutrients<sup>[22]</sup>.

For  $\text{CO}_2$  and  $\text{N}_2\text{O}$  emissions, the GBR model also demonstrated high accuracy (Fig. 3a, b). Both the training and testing datasets closely followed the 1:1 line, with only a 4% difference in  $R^2$  values. The low MSE values further indicate the strong performance of the GBR model in predicting  $\text{CO}_2$  and  $\text{N}_2\text{O}$  emissions under MPs pollution.

The model achieved high  $R^2$  values for both the training and testing datasets on plant biomass (Fig. 4). There was a close alignment between the predicted and observed data along the 1:1 line, with low MSE values, indicating the high performance of the GBR model in predicting plant biomass. These results confirm that the GBR model is an effective tool for predicting, even with small datasets<sup>[23]</sup>.

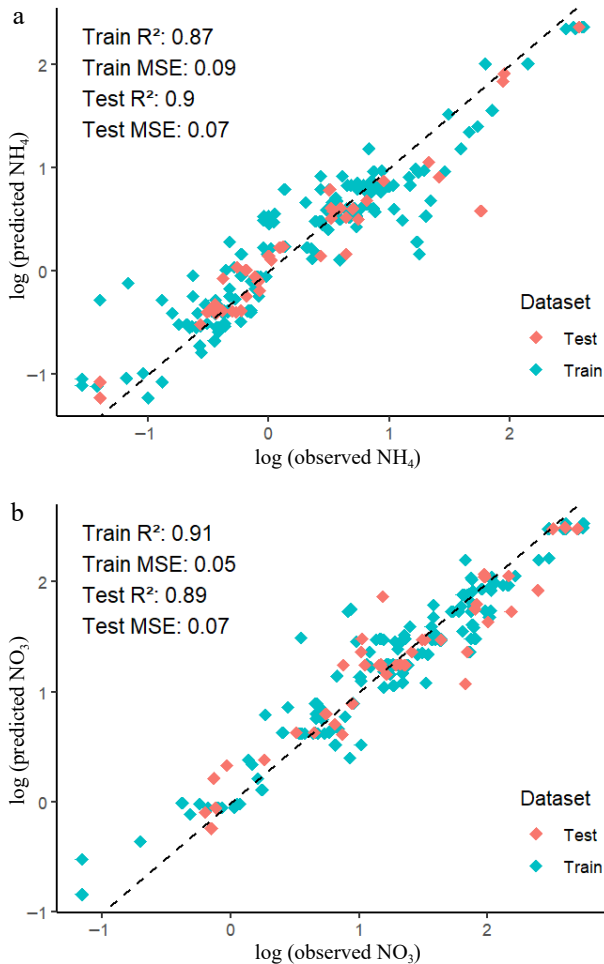
### Contribution of MPs properties, incubation time, and soil type

The spider chart illustrates that soil carbon content varies across experimental conditions (Fig. 5a). MP size accounted for the largest change in soil DOC content (38%), while soil type, incubation time, and MP dose contributed uniformly low changes. Given that the small size of MPs provides a large surface-to-volume ratio, their impact on DOC may result from the physical and chemical changes they induce in the soil<sup>[9]</sup>. Our results are consistent with the findings of Guo et al.<sup>[24]</sup> who found that MP increases DOC contents depending on its size.

Furthermore, we found that MP shape had the greatest impact on SOC content (39%), followed by MP dose (Fig. 5a). In contrast, MP type, MP size, and incubation time contributed minimally to changes in SOC. These findings could be explained by the disintegration of soil aggregates, impacting the stored SOC in soil aggregates<sup>[25,26]</sup>. Typically, the presence of MPs can also affect SOM content, with the magnitude of these effects largely determined by the properties of the MPs<sup>[27,28]</sup>. Thus, our analysis indicates that the properties of MPs contributed distinctly to changes in SOM content (Fig. 5a). However, the most substantial change in SOM content (60%) was attributed to MP shape, suggesting that the diverse shapes of MPs can significantly influence SOM content. The strong influence of microplastic (MP) shape on both SOC and SOM suggests that these variables remain closely related, which is expected given that SOM is commonly estimated from SOC using an empirical conversion factor.

Across experimental conditions, soil  $\text{NH}_4^+$  and  $\text{NO}_3^-$  contents were substantially altered by MP contamination (Fig. 5b). The greatest changes in soil  $\text{NH}_4^+$  (36%) and  $\text{NO}_3^-$  (51%) were caused by MP type, while MP dose and soil type resulted in the smallest changes in soil  $\text{NH}_4^+$ . This corroborates the findings of Greenfield et al.<sup>[29]</sup> and Li et al.<sup>[30]</sup> who found that MP type could substantially affect soil  $\text{NH}_4^+$  and  $\text{NO}_3^-$  content. Considering the diverse types of MPs in soil, we could assume that the reduction of soil N fertility and crop yields is becoming inevitable.

The accumulation of MPs in soil could contribute to climate change by increasing the emissions of greenhouse gases<sup>[31]</sup>. Thus, we observed that greenhouse gas emissions were significantly influenced by MPs' properties and incubation period (Fig. 5c). However, MP size and incubation period led to the highest  $\text{CO}_2$  emissions (34%–39%), while other factors contributed relatively little to  $\text{CO}_2$



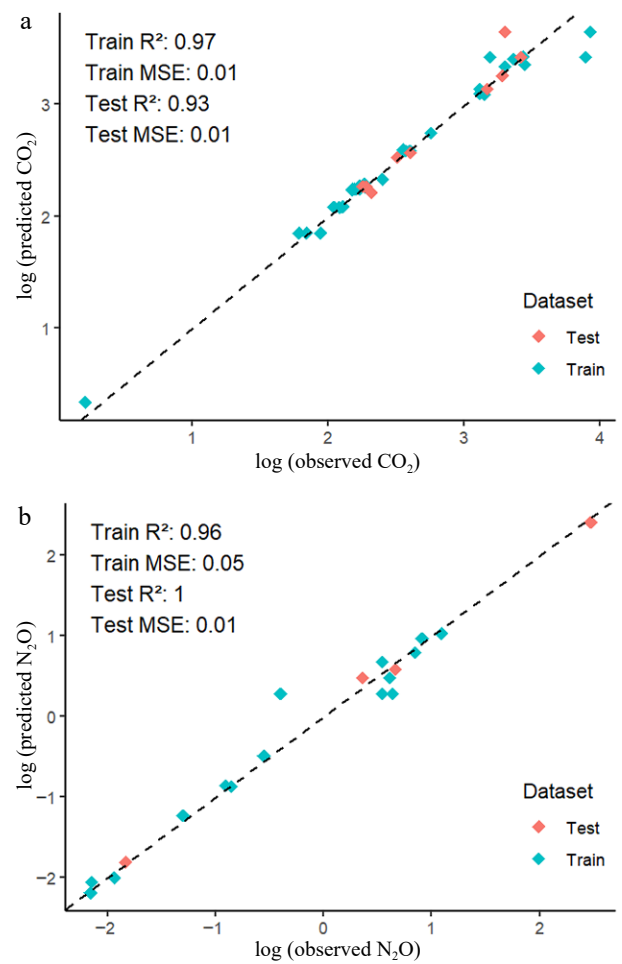
**Fig. 2** Observed vs predicted content of (a) ammonium ( $\text{NH}_4$ ), and (b) nitrate ( $\text{NO}_3$ ) in microplastic polluted soils.

release. These results are consistent with the findings of Iqbal et al.<sup>[2,31]</sup> who found that MP size could substantially affect greenhouse gas emissions. In the case of  $\text{N}_2\text{O}$  emissions, size, type, dose, and incubation period of MPs had substantial effects, whereas soil type had the least impact, contributing only 1%. We can infer from these results that the impacts of MP on  $\text{N}_2\text{O}$  emissions are complex and determined by multiple factors related to pollution. Given the increasing accumulation of MPs in global soils, their impact on climate change is likely to be inevitable.

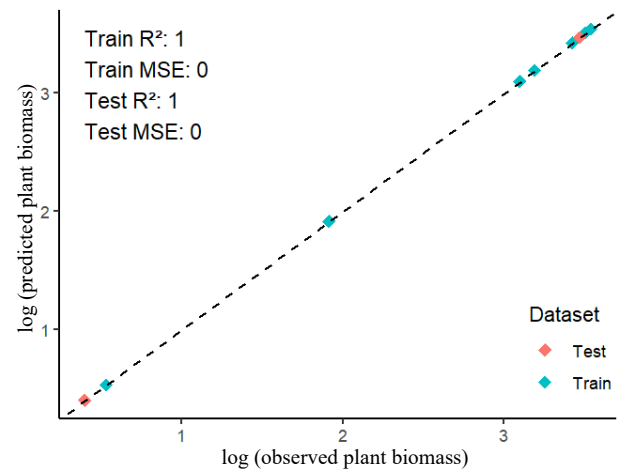
Plant biomass was significantly affected by MP type, size, shape, and dose (Fig. 5d), while incubation time, and soil type had no significant effect. This likely relates to our observed impacts of MPs on C and N cycling, ultimately reducing nutrient availability and altering crop biomass. The greatest change in plant biomass (77%) was attributed to the type of MPs. Our results are consistent with the findings of Wang et al.<sup>[32]</sup> who found that MPs type with Cd or without Cd remarkably influence plant biomass.

## Conclusions

This study demonstrates that the GBR model offers a robust framework for estimating soil C and N cycling under the influence of complex MP properties. However, the robustness of the model for predicting SOC and SOM contents could not be confirmed due to potential model overfitting. Therefore, future studies with larger sample sizes are required to further validate these findings.

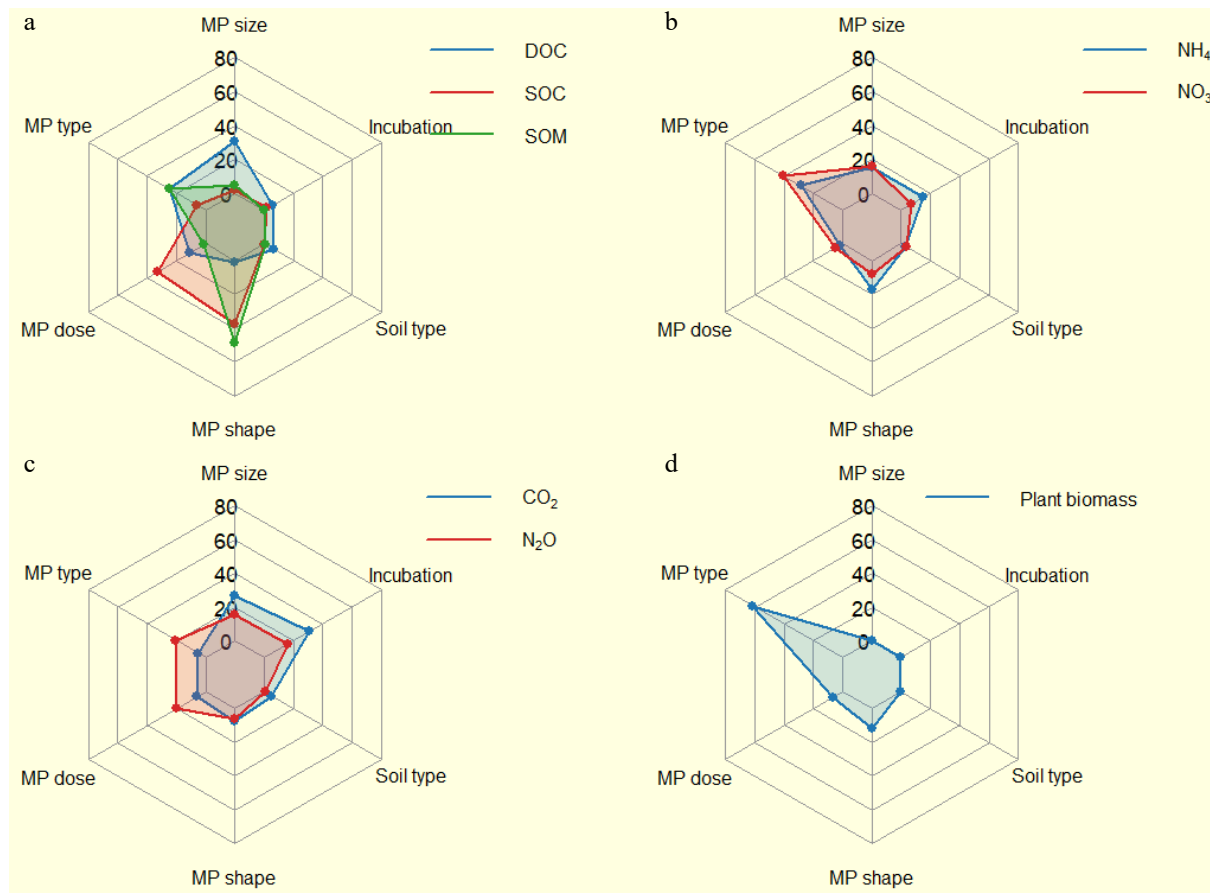


**Fig. 3** Observed vs predicted content of (a) carbon dioxide ( $\text{CO}_2$ ), and (b) nitrous oxide ( $\text{N}_2\text{O}$ ) in microplastic polluted soils.



**Fig. 4** Observed vs predicted effects of soil microplastic pollution on plant biomass.

The GBR simulations reveal that specific MP characteristics exert distinct effects on soil C and N dynamics. The size of MPs can strongly influence DOC and  $\text{CO}_2$  emissions, whereas  $\text{NH}_4^+$  and  $\text{NO}_3^-$  contents are more strongly affected by microplastic type. Given the increasing prevalence of MPs pollution with diverse properties, their impacts on soil C and N cycling are likely to intensify. Under such scenarios, machine learning models offer valuable tools for



**Fig. 5** Contribution of (a) microplastic (MP) properties (type, shape, and size), MP dose, and incubation period, and soil type to soil carbon (SOC, DOC and SOM), (b) nitrogen (NH<sub>4</sub> and NO<sub>3</sub>), (c) greenhouse gas (CO<sub>2</sub> and N<sub>2</sub>O) emissions, and (d) plant biomass. DOC-dissolved organic carbon. SOC-soil organic carbon. SOM-soil organic matter. NH<sub>4</sub>-ammonium. NO<sub>3</sub>- nitrate. CO<sub>2</sub>-carbon dioxide. N<sub>2</sub>O-nitrous oxide.

evaluating ecosystem-level risks and supporting informed decision-making. Importantly, mitigation strategies should move beyond assessments based solely on total MP abundance and instead incorporate specific MP characteristics to better capture their ecological impacts.

### Author contributions

The authors confirm contribution to the paper as follows: study conception and design: Iqbal S, Worthy FR; methodology: Worthy FR; formal analysis: Gui H; data curation: Iqbal S, Gui H; visualization: Gui H; draft manuscript preparation: Iqbal S; writing – review & editing: Iqbal S, Worthy FR, Li Y; supervision: Li Y; funding acquisition: Iqbal S, Li Y. All authors reviewed the results and approved the final version of the manuscript.

### Data availability

The datasets generated during and/or analyzed in the current study are available from the corresponding author on reasonable request.

### Acknowledgments

Authors acknowledge support provided by Zhihui Yunnan Grant (No 202503AM140006) and Yunnan Province Xingdian Talent Support Program.

### Conflict of interest

The authors declare that they have no conflict of interest.

### Dates

Received 19 August 2025; Revised 15 February 2026; Accepted 15 February 2026; Published online 23 March 2026

### References

- [1] Withana PA, Li J, Senadheera SS, Fan C, Wang Y, et al. 2024. Machine learning prediction and interpretation of the impact of microplastics on soil properties. *Environmental Pollution* 341:122833
- [2] Iqbal S, Li Y, Xu J, Worthy FR, Gui H, et al. 2025. Smallest microplastics intensify maize yield decline, soil processes and consequent global warming potential. *Journal of Hazardous Materials* 486:136993
- [3] Iqbal S, Xu J, Gui H, Bu D, Alharbi SA, et al. 2024. Interactive effects of microplastics and typical pollutants on the soil-plant system: a mini-review. *Circular Agricultural Systems* 4:e007
- [4] Iqbal S, Xu J, Khan S, Arif MS, Yasmeen T, et al. 2021. Deciphering microplastic ecotoxicology: impacts on crops and soil ecosystem functions. *Circular Agricultural Systems* 1:1–7
- [5] Iqbal S, Xu J, Allen SD, Khan S, Nadir S, et al. 2020. Unraveling consequences of soil micro- and nano-plastic pollution on soil-plant system: Implications for nitrogen (N) cycling and soil microbial activity. *Chemosphere* 260:127578

- [6] Luo Y, Wang L, Cao T, Chen J, Lv M, et al. 2023. Microplastics are transferred by soil fauna and regulate soil function as material carriers. *Science of the Total Environment* 857:159690
- [7] Maddela NR, Ramakrishnan B, Kadiyala T, Venkateswarlu K, Megharaj M. 2023. Do microplastics and nanoplastics pose risks to biota in agricultural ecosystems? *Soil Systems* 7(1):19
- [8] Kim K, Song IG, Yoon H, Park JW. 2023. Sub-micron microplastics affect nitrogen cycling by altering microbial abundance and activities in a soil-legume system. *Journal of Hazardous Materials* 460:132504
- [9] Iqbal S, Xu J, Arif MS, Worthy FR, Jones DL, et al. 2024. Do added microplastics, native soil properties, and prevailing climatic conditions have consequences for carbon and nitrogen contents in soil? a global data synthesis of pot and greenhouse studies. *Environmental Science & Technology* 58:8464–8479
- [10] Ma Y, Yang K, Yu H, Tan W, Gao Y, et al. 2024. Effects and mechanism of microplastics on organic carbon and nitrogen cycling in agricultural soil: a review. *Soil Use and Management* 40:e12971
- [11] Seeley ME, Song B, Passie R, Hale RC. 2020. Microplastics affect sedimentary microbial communities and nitrogen cycling. *Nature Communications* 11:2372
- [12] Zhao T, Lozano YM, Rillig MC. 2021. Microplastics increase soil pH and decrease microbial activities as a function of microplastic shape, polymer type, and exposure time. *Frontiers in Environmental Science* 9:675803
- [13] Zhu J, Liu S, Wang H, Wang D, Zhu Y, et al. 2022. Microplastic particles alter wheat rhizosphere soil microbial community composition and function. *Journal of Hazardous Materials* 436:129176
- [14] Lozano YM, Aguilar-Trigueros CA, Onandia G, Maaß S, Zhao T, et al. 2021. Effects of microplastics and drought on soil ecosystem functions and multifunctionality. *Journal of Applied Ecology* 58:988–996
- [15] Shen M, Song B, Zhou C, Almatrafi E, Hu T, et al. 2022. Recent advances in impacts of microplastics on nitrogen cycling in the environment: a review. *Science of the Total Environment* 815:152740
- [16] Lozano YM, Caesaria PU, Rillig MC. 2022. Microplastics of different shapes increase seed germination synchrony while only films and fibers affect seed germination velocity. *Frontiers in Environmental Science* 10:1017349
- [17] Yuan X, Suvarna M, Low S, Dissanayake PD, Lee KB, et al. 2021. Applied machine learning for prediction of CO<sub>2</sub> adsorption on biomass waste-derived porous carbons. *Environmental Science & Technology* 55:11925–11936
- [18] Shi L, Li J, Palansooriya KN, Chen Y, Hou D, et al. 2023. Modeling phytoremediation of heavy metal contaminated soils through machine learning. *Journal of Hazardous Materials* 441:129904
- [19] Li J, Li L, Tong YW, Wang X. 2023. Understanding and optimizing the gasification of biomass waste with machine learning. *Green Chemical Engineering* 4:123–133
- [20] Li J, Pan L, Suvarna M, Tong YW, Wang X. 2020. Fuel properties of hydrochar and pyrochar: prediction and exploration with machine learning. *Applied Energy* 269:115166
- [21] Xia Y, Sun J, Chen DG. 2018. Introduction to R, RStudio and ggplot2. In *Statistical analysis of microbiome data with R*. Singapore: Springer. pp. 77–127 doi: 10.1007/978-981-13-1534-3\_4
- [22] Gökmen F, Uygur V, Sukuşu E. 2023. Extreme gradient boosting regression model for soil available boron. *Eurasian Soil Science* 56:738–746
- [23] Chen T, Guestrin C. 2016. XGBoost: a scalable tree boosting system. *Kokusai Denshin Denwa Company (KDD)' 16: Proceedings of the 22<sup>nd</sup> ACM SIGKDD International Conference on Knowledge Discovery and Data Mining, August 13–17, 2016, San Francisco, CA, USA*. pp. 785–794 doi: 10.1145/2939672.2939785
- [24] Guo Z, Li P, Ma L, Yang X, Yang J, et al. 2024. Cascading effects from soil to maize functional traits explain maize response to microplastics disturbance in multi-nutrient soil environment. *Geoderma* 441:116759
- [25] Lehmann A, Leifheit EF, Gerdawischke M, Rillig MC. 2021. Microplastics have shape- and polymer-dependent effects on soil processes. *Microplastics and Nanoplastics* 1:1–14
- [26] Zhang GS, Liu YF. 2018. The distribution of microplastics in soil aggregate fractions in southwestern China. *Science of the Total Environment* 642:12–20
- [27] Dong Y, Gao M, Qiu W, Song Z. 2021. Effect of microplastics and arsenic on nutrients and microorganisms in rice rhizosphere soil. *Ecotoxicology and Environmental Safety* 211:111899
- [28] Boots B, Russell CW, Green DS. 2019. Effects of microplastics in soil ecosystems: above and below ground. *Environmental Science & Technology* 53:11496–11506
- [29] Greenfield LM, Graf M, Rengaraj S, Bargiela R, Williams G, et al. 2022. Field response of N<sub>2</sub>O emissions, microbial communities, soil biochemical processes and winter barley growth to the addition of conventional and biodegradable microplastics. *Agriculture, Ecosystems & Environmental* 336:108023
- [30] Li C, Cui Q, Li Y, Zhang K, Lu X, et al. 2022. Effect of LDPE and biodegradable PBAT primary microplastics on bacterial community after four months of soil incubation. *Journal of Hazardous Materials* 429:128353
- [31] Iqbal S, Xu J, Saleem Arif M, Shakoor A, Worthy FR, et al. 2024. Could soil microplastic pollution exacerbate climate change? A meta-analysis of greenhouse gas emissions and global warming potential. *Environmental Research* 252:118945
- [32] Wang F, Zhang X, Zhang S, Zhang S, Sun Y. 2020. Interactions of microplastics and cadmium on plant growth and arbuscular mycorrhizal fungal communities in an agricultural soil. *Chemosphere* 254:126791



Copyright: © 2026 by the author(s). Published by Maximum Academic Press, Fayetteville, GA. This article is an open access article distributed under Creative Commons Attribution License (CC BY 4.0), visit <https://creativecommons.org/licenses/by/4.0/>.