

Original Research

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Crystalline phase-dependent hydrolysis of organophosphate esters by iron oxyhydroxides: implications for nanomineral-mediated transformation of plastic additives

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Abstract

A key risk factor associated with pervasive plastic pollution is the release of additives, including organophosphate esters (OPEs). Metal-bearing nanoparticles, such as iron oxides, can significantly affect the transformation of these plastic additives. However, the mechanisms underlying how intrinsic properties of iron oxide minerals determine their capability to mediate the transformation of OPEs are still obscure. Here, it is demonstrated that iron oxyhydroxide nanoparticles (goethite, akaganeite, and lepidocrocite) can effectively catalyze the hydrolysis of 4-nitrophenyl phosphate (pNPP), a model OPE contaminant, under environmentally relevant pH conditions (pH 6.0 to 8.0). The catalytic efficiency exhibits a pronounced dependence on the crystalline phase of the iron oxyhydroxides, with surface area-normalized rate constant following the order: lepidocrocite > akaganeite > goethite. The crystalline phase-dependent catalytic performance is jointly governed by two key factors, i.e., the adsorption affinity toward pNPP of the iron oxyhydroxide nanoparticles and the reactivity of adsorbed pNPP at the surface. Specifically, differences in surface charge distribution among the crystalline phases lead to variations in electronic interactions with the phosphate groups in pNPP, affecting their adsorption affinities (akaganeite > goethite > lepidocrocite). Concurrently, the intrinsic catalytic reactivity (lepidocrocite > goethite > akaganeite) is governed by variations in Lewis acid sites and their ability to induce charge redistribution at the central P atom in adsorbed pNPP. These findings underscore the critical role of the mineral crystalline phase in controlling the environmental fate of OPEs and other plastic-derived additives, and can guide the accurate evaluation of the ecological risks of these emerging pollutants.

Keywords: Organophosphate esters, Iron oxyhydroxide, Crystalline phase, Hydrolysis, Plastic pollution

Highlights

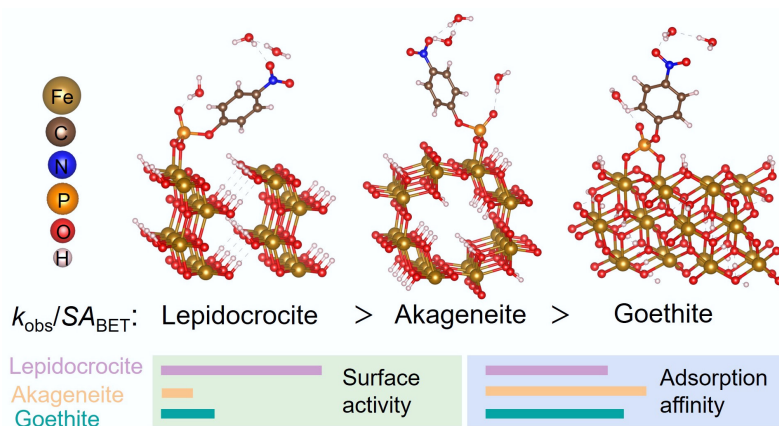
- Crystalline phase dictates the efficiency of iron oxyhydroxides in catalyzing OPE hydrolysis.
- Lepidocrocite exhibits superior catalytic performance to goethite and akaganeite.
- Crystalline phase-dependent Fe 3d-O 2p electronic interaction mediates adsorption affinity.
- Crystalline phase-dependent Lewis acidity determines the surface reactivity of adsorbed OPEs.

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Graphical abstract



Introduction

The increasing prevalence of plastic pollution has garnered world-wide attention^[1–3], and one key risk factor associated with plastics pollution is the release of additives^[4], such as phthalates and organophosphate esters (OPEs)^[5–7]. In addition to being a common plasticizer, OPEs are also extensively utilized as pesticides and flame retardants^[8–10], and are commonly detected in the environment^[11–14]. The majority of OPE pollutants are endocrine disruptors^[15–18], and can induce a wide range of toxic effects, including neurological^[19], respiratory^[20], developmental^[21–23], and reproductive^[24] toxicities, as well as potential carcinogenicity^[25–27], thus causing concerns over their health and ecological risks. The degradation and weathering of plastics into microscale and nanoscale particles is expected to accelerate the leaching of OPE additives^[28–30], which exacerbates the environmental impacts of microplastics^[28,31,32]. The ecological risks of OPE pollutants are largely dictated by their environmental behavior and transformation processes, including biotic and abiotic oxidation and hydrolysis^[33–35].

Metal-bearing nanoparticles, which are abundant in the environment, play significant roles in various biogeochemical processes and substantially impact the fate and transport of environmental pollutants^[36–38]. Iron oxides, among the most plentiful metal oxide nanoparticles in the environment^[36], can significantly influence the transformation of pollutants^[39,40], due to their compositional and structural diversity and redox activity^[39,41]. For instance, nanosized iron oxides can adsorb pollutants^[42,43], or form co-precipitates with them^[44]. Moreover, iron oxides can catalyze the oxidation-reduction^[45–47] and hydrolysis^[34,48–50] reactions of pollutants in the environment. Hydrolysis is one of the most important reaction types that affect the environmental fate of organic pollutants such as esters^[48,51], chlorinated hydrocarbons^[52], and antibiotics^[50,53,54]. Recent studies have demonstrated that physicochemical properties of nanosized iron oxides, particularly exposed facets, can affect their efficiency in mediating the hydrolytic transformation of organic pollutants (including phthalates and OPEs) in the environment^[48,51]. However, the impacts of crystalline phase, a critical parameter that fundamentally determines the bulk and surface atomic arrangement and coordination state, on the ability of iron oxide nanoparticles to mediate the hydrolysis of organic pollutants remain to be elucidated.

Herein, we investigated the mechanisms by which the crystalline phase of iron oxyhydroxide nanoparticles affects their efficiencies in

mediating the hydrolysis of OPE pollutants. Three common polymorphs of iron oxyhydroxide, including goethite (α -FeOOH), akageneite (β -FeOOH), and lepidocrocite (γ -FeOOH), were chosen as model materials, and p-nitrophenyl phosphate (pNPP) was selected as a model OPE compound, as it represents the core structure of common OPE pollutants^[55,56]. The performance of the oxyhydroxide nanoparticles in catalyzing pNPP hydrolysis at environmentally relevant pH (6.0 to 8.0) was examined, and the Langmuir–Hinshelwood kinetics model was used to compare the differences in pNPP adsorption affinity and surface catalytic activity among the iron oxyhydroxide nanoparticles. Moreover, the adsorption configuration of pNPP and its adsorption affinity on iron oxyhydroxide with different crystalline phases were investigated using *in situ* attenuated total reflectance Fourier transform infrared spectroscopy (ATR-FTIR) analysis, and density functional theory (DFT) calculations. A correlation between surface reactivity, and Lewis acidity of the materials was discerned using Fourier transform infrared spectroscopy after pyridine adsorption (Py-IR), and DFT calculations. This study provides valuable insights to enhance predictive understanding of the environmental behavior of OPE plastic additives and the related ecological risks.

Methods

Synthesis and characterization of materials

Information on the chemicals used is provided in [Supplementary File 1](#) (Text S1). The specific synthesis procedures for the three iron oxyhydroxide materials are detailed in [Supplementary File 1](#) (Text S2).

The morphological features of the materials were characterized using scanning electron microscopy (SEM) and transmission electron microscopy (TEM). The crystalline phases were determined using X-ray diffraction (XRD). The specific surface area (SA_{BET}) was determined from the nitrogen adsorption/desorption isotherms. The ζ -potential was measured at pH 3 to 11. The surface acidity of the materials was analyzed by Py-IR^[57]. More detailed characterization methods are provided in [Supplementary File 1](#) (Text S3).

Hydrolysis experiments

The pNPP hydrolysis experiments were performed at environmentally relevant conditions (pH 6.0 to 8.0, 25 ± 0.1 °C). The concentrations of pNPP and 4-nitrophenol (4-NP) over time were determined from the UV-vis absorbance at 310 and 400 nm, respectively, with their UV-vis spectra shown in [Supplementary Fig. S1](#). Experiments were carried out

in triplicate. More detailed methods are provided in [Supplementary File 1](#) (Text S4).

Phosphate adsorption experiments

Since pNPP constantly hydrolyzes, which makes it inaccurate to directly determine the adsorption capacity of the iron oxyhydroxide nanoparticles for pNPP, the adsorption of orthophosphate ions as a surrogate was measured^[58]. The adsorption isotherms were determined at pH 6.0, 7.0, and 8.0 (with detailed methods in [Supplementary File 1](#), Text S5).

Data analysis

The pNPP hydrolysis data were fitted to the pseudo-first-order kinetic model to calculate the apparent rate constant, k_{obs} (h^{-1}).

The Langmuir–Hinshelwood (L–H) model, a widely used model to describe the kinetics of heterogeneous catalytic reactions^[59], was employed to obtain the adsorption coefficient of pNPP on the materials, K_L (L mg^{-1}) and the reaction rate constant of adsorbed pNPP, k_r ($\text{mg L}^{-1} \text{h}^{-1}$). More details are provided in [Supplementary File 1](#) (Text S6).

In situ ATR-FTIR measurements of pNPP adsorption

In situ ATR-FTIR spectroscopy was employed to investigate the adsorption configuration of pNPP on the surface of iron oxyhydroxides. Detailed methodologies are provided in [Supplementary File 1](#) (Text S7).

Theoretical calculations

DFT calculations were carried out using the VASP 5.4.4 package^[60], to obtain information on the electron localization function (ELF), pNPP adsorption energy, charge density differences, partial density of states (PDOS), and crystal orbital Hamilton population (COHP). Details are provided in [Supplementary File 1](#) (Text S8).

Statistical analysis

Statistical differences were tested by one-way analysis of variance (ANOVA), using Statistical Product and Service Solutions (version 19.0, IBM, USA), with significant differences ($p < 0.05$) indicated by different letters. For pNPP hydrolysis and phosphate ion adsorption experiments, arithmetic means, and standard deviations of triplicate tests were reported.

Results and discussion

Physicochemical properties of iron oxyhydroxide nanoparticles

The as-synthesized iron oxyhydroxide materials have different crystalline phases, as demonstrated by XRD results. The XRD pattern of goethite (JCPDS No. 29–0713)^[61] has diffraction peaks at $2\theta = 17.8^\circ$, 21.2° , 33.2° , 34.7° , 36.6° , and 53.2° , corresponding to the (020), (110), (130), (021), (111), and (221) crystal planes, respectively. The XRD pattern of akaganeite (JCPDS No. 34-1266)^[61] has diffraction peaks at $2\theta = 11.8^\circ$, 16.7° , 26.6° , 33.9° , 35.1° , 39.2° , 46.4° , and 55.8° , corresponding to the (110), (200), (310), (400), (211), (301), (411), and (521) crystal planes, respectively. The XRD pattern of lepidocrocite (JCPDS No. 08-0098)^[62] has diffraction peaks at $2\theta = 14.2^\circ$, 27.1° , 36.4° , and 46.8° , corresponding to the (020), (120), (031), and (200) crystal planes, respectively ([Fig. 1a](#)).

Nitrogen adsorption-desorption isotherm analyses ([Fig. 1b](#)) revealed that lepidocrocite possesses a higher S_{BET} ($169.2 \text{ m}^2 \text{ g}^{-1}$)

than goethite ($71.9 \text{ m}^2 \text{ g}^{-1}$), and akaganeite ($45.8 \text{ m}^2 \text{ g}^{-1}$)^[63]. The larger specific surface area, which is conducive to exposure of more catalytic sites, likely can improve the catalytic efficiency. The isotherms are type IV with type H3 hysteresis loops, indicating the presence of mesopores and macropores in the aggregates^[63,64], which is consistent with pore size distribution analysis results ([Fig. 1c](#)). Among the three materials, lepidocrocite also exhibits the narrowest pore size distribution, which may hinder the diffusion, and mass transfer of target molecules.

The SEM and TEM images of three types of iron oxyhydroxide nanoparticles ([Fig. 1d–f](#)) show that goethite exhibits nanorod morphology with lengths of 200–600 nm and diameters of 40–50 nm, with most nanorods forming bundled aggregates ([Fig. 1d](#)). In contrast, akaganeite presents more uniform spindle-like structures ranging from 200–600 nm in length, and 50–150 nm in diameter ([Fig. 1e](#)), and lepidocrocite is composed of nanofibers with diameters of 10–30 nm ([Fig. 1f](#)). Furthermore, in the high-resolution TEM (HRTEM) images, the goethite revealed lattice spacing of 0.26 nm, indexed to the (021) plane^[65], and the akaganeite and lepidocrocite displayed lattice spacing of 0.27 and 0.33 nm, corresponding to the (400) plane and (200) plane^[66], respectively. Additionally, structural modeling and surface ELF analysis ([Fig. 1g–i](#)) reveal distinct densities of unsaturated Fe sites per unit surface area across the different FeOOH phases. This crystalline phase dependency originates not only from active site density differences but also from coordination environments and electronic structures of surface Fe atoms, which collectively modulate Lewis acidity and catalytic reactivity^[48,67].

The crystalline phase influences the catalytic pNPP hydrolysis performance of iron oxyhydroxide nanoparticles

The crystalline phase plays a key role in determining the activity of the iron oxyhydroxide nanoparticles in catalyzing the hydrolysis of pNPP under environmentally relevant pH conditions (6.0, 7.0, and 8.0). All three iron oxyhydroxide nanoparticles (100 mg L^{-1}) significantly accelerated pNPP hydrolysis ([Fig. 2](#)), with lepidocrocite exhibiting the highest activity across all pH levels. For instance, at pH 6.0, the removal efficiency of pNPP reached 85% (lepidocrocite), 45% (goethite), and 25% (akaganeite) after 48 h. It is important to note that the mass balance of pNPP and its hydrolysis reaction product 4-NP in the aqueous phase remained > 95% ([Supplementary Figs S2–S4](#)). Considering that pNPP hydrolysis was negligible (< 3%) in the absence of iron oxyhydroxide nanoparticles ([Fig. 2a](#)), we believe that heterogeneous catalytic hydrolysis was the dominant removal mechanism rather than adsorption. The hydrolysis kinetics fitted to the pseudo-first-order model well, and surface area-normalized rate constant ($k_{\text{obs}}/S_{\text{BET}}$) followed the order of lepidocrocite > akaganeite > goethite ([Fig. 2d–f](#) and [Supplementary Table S1](#)). Additionally, hydrolysis efficiency decreased with increasing pH, which can be attributed to electrostatic repulsion between the negatively charged iron oxyhydroxides surfaces at higher pH ([Supplementary Fig. S5](#)) and the deprotonated forms of pNPP (pNPP^- and/or pNPP^{2-}) predominant at pH 6.0 to 8.0 ([Supplementary Fig. S6](#)), thereby hindering substrate adsorption and diminishing catalytic performance^[68].

The hydrolysis of pNPP catalyzed by iron oxyhydroxide nanoparticles proceeds via a surface-mediated reaction, in which pNPP first adsorbs to the surface active sites, followed by cleavage of the phosphate ester bond. To decouple the effects of adsorption and surface reactivity, initial reaction rates (r_0) were fitted using the L–H model^[59] across variable initial pNPP concentrations ($C_0 = 3.2\text{--}16 \text{ mg L}^{-1}$) and pH values ([Supplementary Tables S2–S4](#), [Fig. 3a–f](#), [Supplementary Figs. S7–S9](#)). For all the iron oxyhydroxides, r_0 as a function

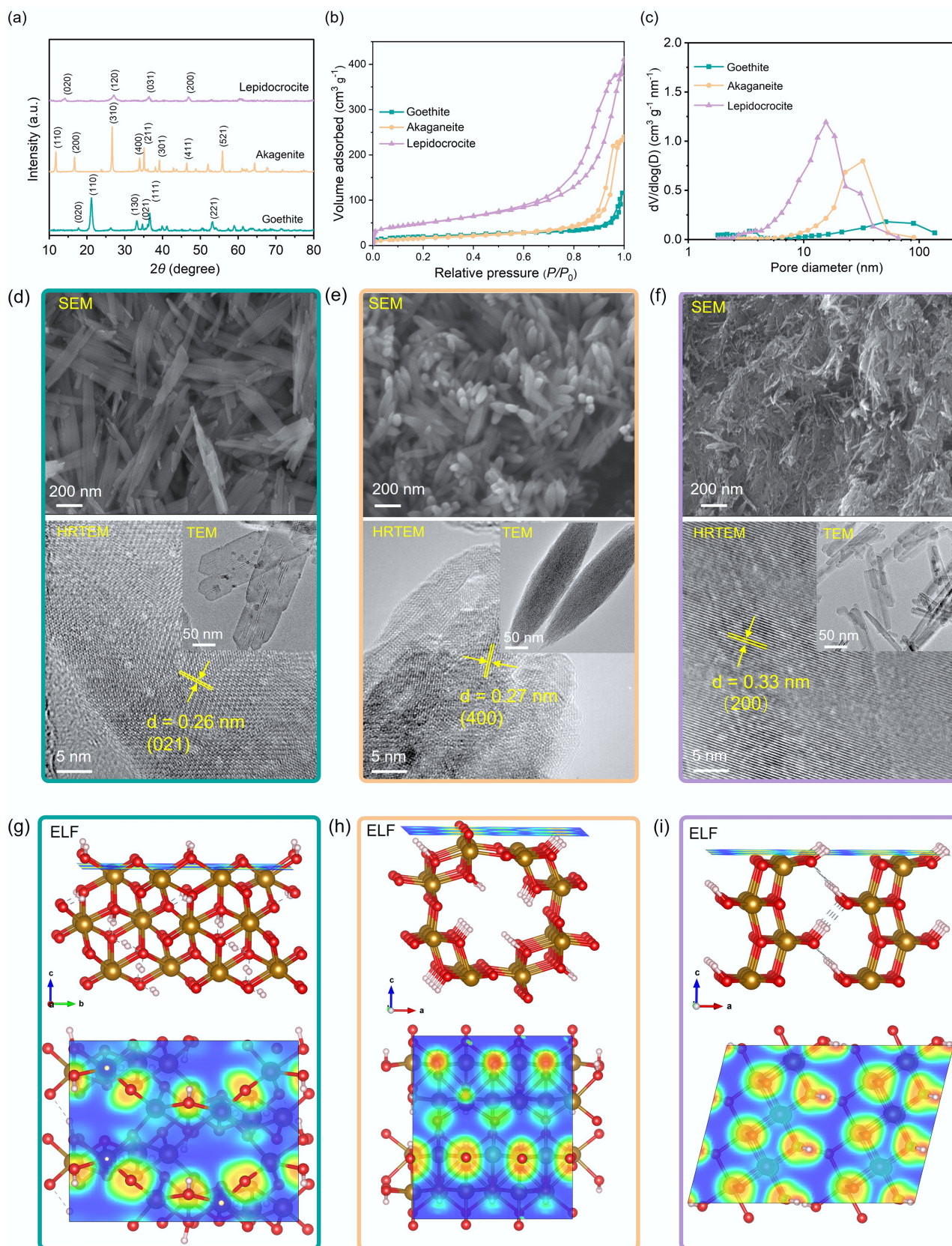


Fig. 1 (a) X-ray diffraction (XRD) patterns, (b) nitrogen adsorption–desorption isotherms, and (c) pore size distribution curves of goethite, akaganeite, and lepidocrocite. Scanning electron microscopy (SEM), transmission electron microscopy (TEM), and high-resolution TEM (HRTEM) images of (d) goethite, (e) akaganeite, and (f) lepidocrocite. Electron localization function (ELF) analysis of (g) goethite, (h) akaganeite, and (i) lepidocrocite. Red regions in the ELF isosurface represent strong electron localization, while blue regions indicate delocalized electron states.

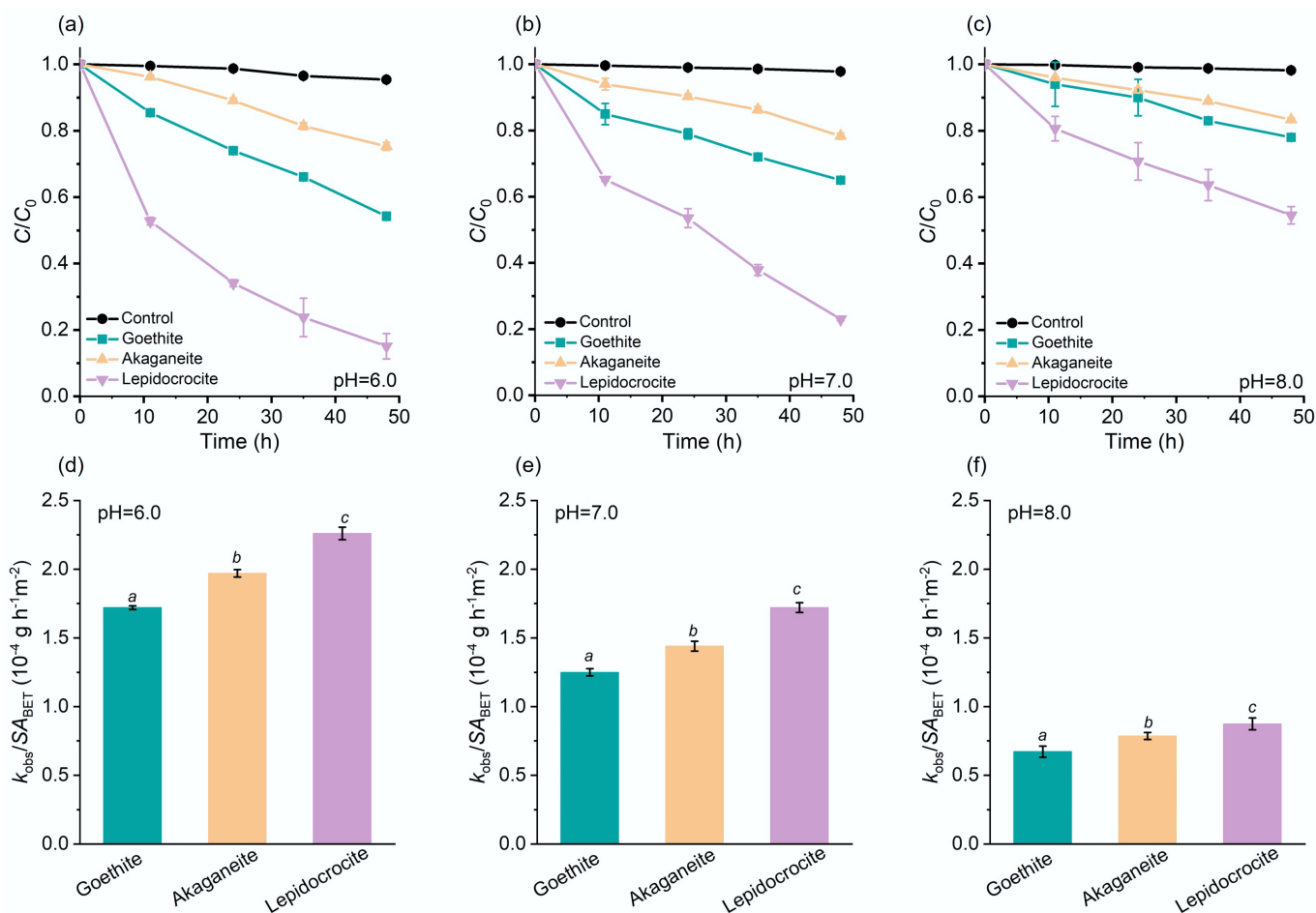


Fig. 2 Efficiencies of iron oxyhydroxide nanoparticles in catalyzing pNPP hydrolysis. Catalytic pNPP hydrolysis kinetic curves in the absence (denoted as 'Control'), or presence of the iron oxyhydroxide nanoparticles at pH (a) 6.0, (b) 7.0, (c) 8.0, and (d)–(f) the corresponding surface area-normalized k_{obs} . Initial pNPP concentration: 6.0 mg L⁻¹. The italic letters denote significant differences ($p < 0.05$).

of C_0 fits well with the L–H model (Fig. 3a–f), revealing distinct differences in both adsorption affinity and surface reactivity (Fig. 3g–i). At pH 6 (Fig. 3g), the K_L values followed the trend: akaganeite (0.685 L mg⁻¹) > goethite (0.591 L mg⁻¹) > lepidocrocite (0.521 L mg⁻¹), indicating the strongest pNPP affinity of akaganeite. Interestingly, this trend does not align with the apparent catalytic activity normalized to surface area ($k_{\text{obs}}/SA_{\text{BET}}$), and notably, lepidocrocite, which exhibited the highest $k_{\text{obs}}/SA_{\text{BET}}$, had the lowest affinity toward pNPP. This suggests that, apart from adsorption affinity, other factors also influence the overall hydrolysis efficiency. Indeed, the surface reaction constants show a different pattern, and lepidocrocite exhibited much higher k_r than the other two materials (Fig. 3j–l); for example, at pH 6, lepidocrocite exhibited a 5.1-fold higher k_r value than akaganeite and 3.0 times greater than goethite (Fig. 3j). The highest surface reactivity of lepidocrocite compensated for its lowest pNPP adsorption affinity, thus contributing to its superior overall catalytic efficiency.

Crystalline phase regulates pNPP adsorption affinity by modulating electronic interaction of Fe 3d and O 2p

The effective adsorption of reactant molecules onto a material surface is a prerequisite for heterogeneous catalytic hydrolysis of organic contaminants mediated by metal oxide nanoparticles^[48,53].

Accordingly, differences in the adsorption affinity of pNPP on iron oxyhydroxides with different crystalline phases have a crucial influence on their overall catalytic performance. *In situ* ATR-FTIR analysis suggests that pNPP molecules were primarily adsorbed through inner-sphere complexation, which involves the $-PO_4$ group of pNPP binding directly to surface Fe atoms. The characteristic absorption bands of pNPP on the material surfaces progressively increased over time (Fig. 4a), reflecting the continuous adsorption process^[69]. Although the ATR-FTIR spectral features differed among goethite, akaganeite, and lepidocrocite, likely due to variations in surface charge distribution caused by different crystalline phases (Fig. 1g–i), several common trends were observed, providing insights into their adsorption mechanisms. Specifically, the ATR-FTIR spectrum of pNPP dissolved in water (Supplementary Fig. S10) showed bands of the $-PO_4$ group at 1,089, 1,167, and 1,260 cm⁻¹^[43,48,69], and that of the $-NO_2$ group at 1,347 cm⁻¹ (corresponding to the $\nu_s(-NO_2)$ mode)^[48,70]. When pNPP was adsorbed onto iron oxyhydroxides (Fig. 4a), the characteristic $\nu(P-O)$, $\nu_s(PO_2^-)$, and $\nu(P=O)$ bands of the $-PO_4$ group increased in intensity over time and shifted to higher wavenumbers, whereas that of the $-NO_2$ group remained stable at 1,345–1,371 cm⁻¹^[43,48]. Moreover, a band at 950–990 cm⁻¹ was observed (Fig. 4a), which corresponds to the P–O–Fe bond^[48,71,72], arising from the inner-sphere complexation between pNPP and the surface Fe atoms of the iron oxyhydroxide nanoparticles.

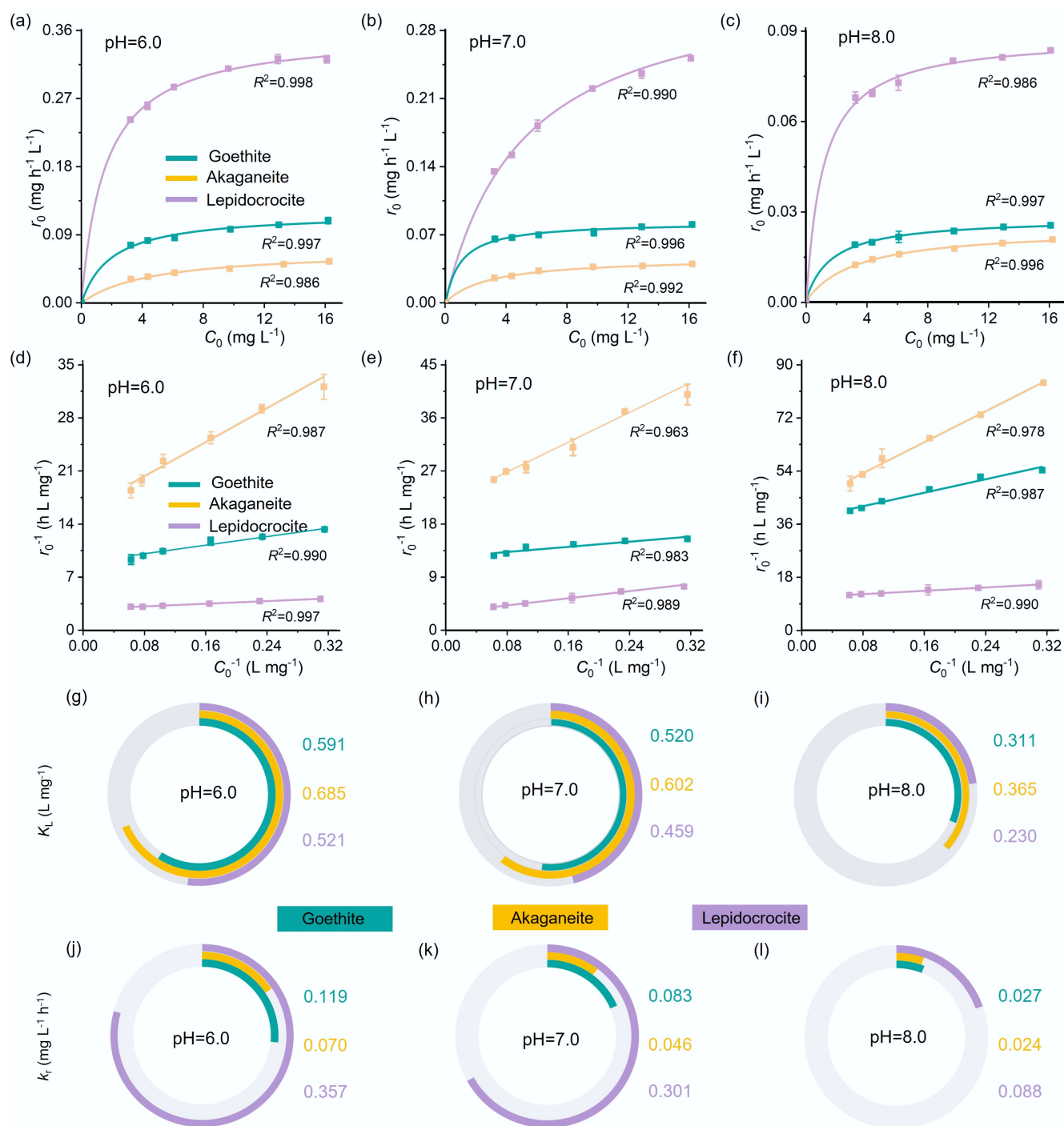


Fig. 3 (a)–(c) Initial hydrolysis reaction rate (r_0) vs initial pNPP concentration (C_0), with lines representing least-squares fit to the L–H model, (d)–(f) linear plot of r_0^{-1} vs C_0^{-1} , (g)–(i) L–H model parameters K_L (representing surface adsorption affinity), and (j)–(l) k_r (representing surface reactivity) determined for each material at pH 6.0, 7.0, and 8.0.

DFT calculations confirmed that the adsorption energy (E_{ads}) of pNPP via the $-\text{PO}_4$ group on iron oxyhydroxide surfaces was considerably larger in absolute value than that through the $-\text{NO}_2$ group (Fig. 4b, Supplementary Figs S11–S16), indicating that adsorption via the $-\text{PO}_4$ group is thermodynamically more favorable. This result corroborates the proposed mechanism that pNPP preferentially binds to surface Fe atoms on iron oxyhydroxides through inner-sphere complexation involving the $-\text{PO}_4$ group. Additionally, noticeable differences in E_{ads} values were observed among three

crystalline phases of iron oxyhydroxides. Specifically, akaganeite exhibited the lowest E_{ads} (–5.44 eV), followed by goethite (–4.86 eV), and lepidocrocite (–4.20 eV), suggesting that akaganeite possesses a stronger adsorption affinity for pNPP, consistent with the experimental observations (Fig. 3g–i). Moreover, the PDOS and COHP analyses revealed a more pronounced overlap between Fe 3d orbitals and O 2p orbitals of the $-\text{PO}_4$ group compared with the $-\text{NO}_2$ group (Fig. 4c–h). The bonding orbitals were more extensively populated with covalent electrons ($-\text{COHP} > 0$, $E < E_{\text{f}}$, $-\text{iCOHP} = 0.027\text{--}0.041$),

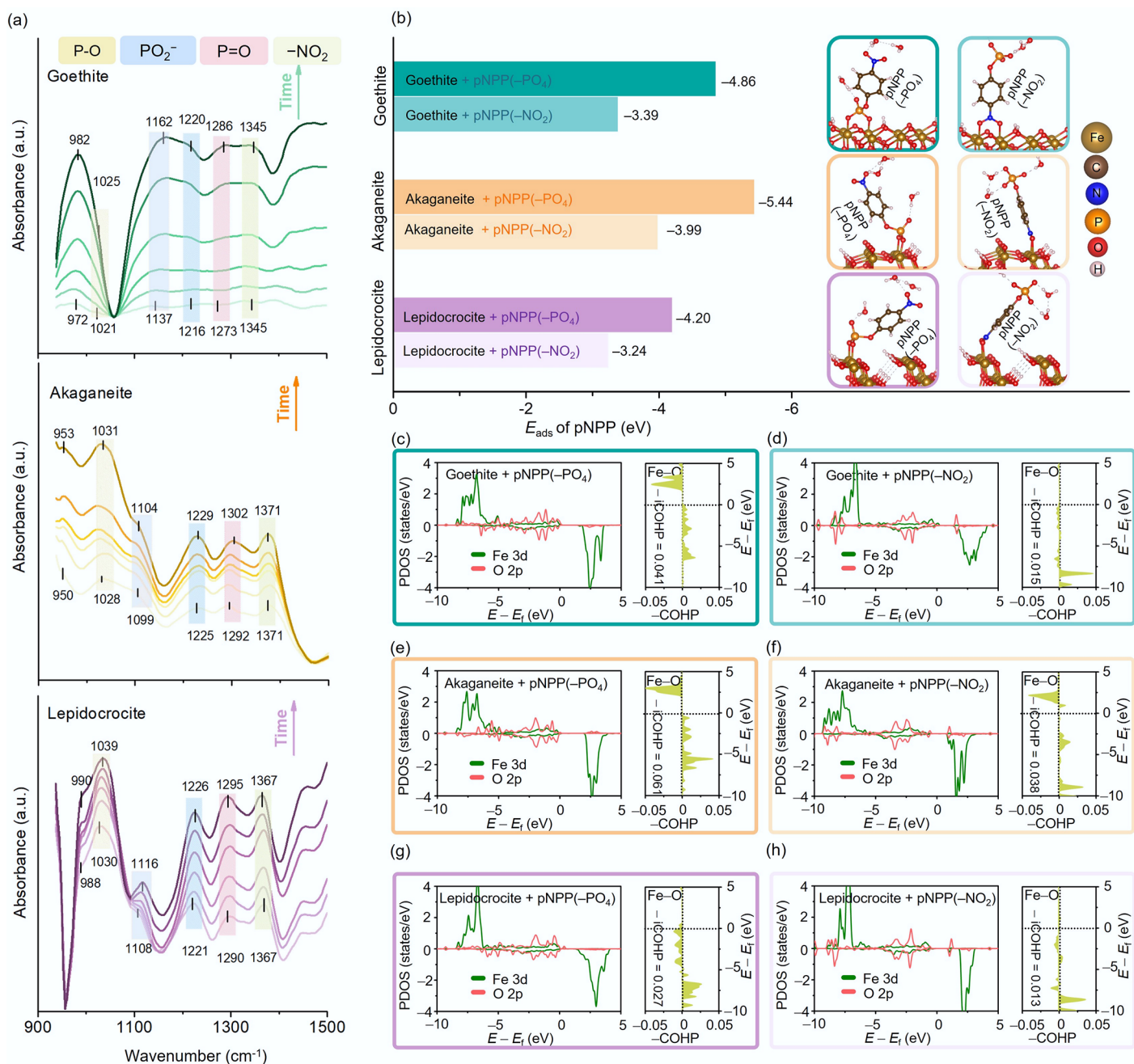


Fig. 4 (a) *In situ* ATR-FTIR spectra of goethite, akaganeite, and lepidocrocite (along the direction indicated by the arrow; the time is: 5, 30, 60, 90, 130, and 180 min). (b) The adsorption energy (E_{ads}) of pNPP on goethite, akaganeite, and lepidocrocite via different functional groups ($-\text{PO}_4$ or $-\text{NO}_2$) and the corresponding stable adsorption configurations (insets). The PDOS and COHP of Fe 3d and O 2p orbitals during adsorption in various configurations: adsorption of pNPP via the $-\text{PO}_4$ or $-\text{NO}_2$ group on (c), (d) goethite, (e), (f) akaganeite, and (g), (h) lepidocrocite.

indicating a stronger electronic interaction between Fe atoms and the $-\text{PO}_4$ group^[73]. Among the crystalline phases, akaganeite displayed the highest bonding strength, as reflected by its $-i\text{COHP}$ value of 0.061, exceeding those of goethite (0.041), and lepidocrocite (0.027). This suggests a higher degree of electron density in the bonding orbitals and stronger interaction between akaganeite and pNPP. Interestingly, surface area-normalized adsorption coefficients K_d of orthophosphate ions by the materials (Supplementary Fig. S17) showed similar trends to the adsorption affinity predicted by theoretical calculations, which is also consistent with the experimentally determined K_L of pNPP.

Crystalline phase enhances catalytic hydrolysis performance by inducing electron redistribution in pNPP via surface Lewis acid sites

The coordination saturation of metal sites on metal oxide surfaces is a key factor governing their Lewis acidity, which directly influences their catalytic activity toward the hydrolysis of organophosphorus pollutants^[34,48,74]. According to Py-IR analysis (Fig. 5a and Supplementary Fig. S18), all three types of iron oxyhydroxides are rich in Lewis acid sites (the bands at 1,448, 1,597, and 1,608 cm^{-1} are typically attributed to Lewis acid sites^[47,75]; that at 1,492 cm^{-1} is usually assigned to both Lewis and Brønsted acid sites^[75], though with weaker

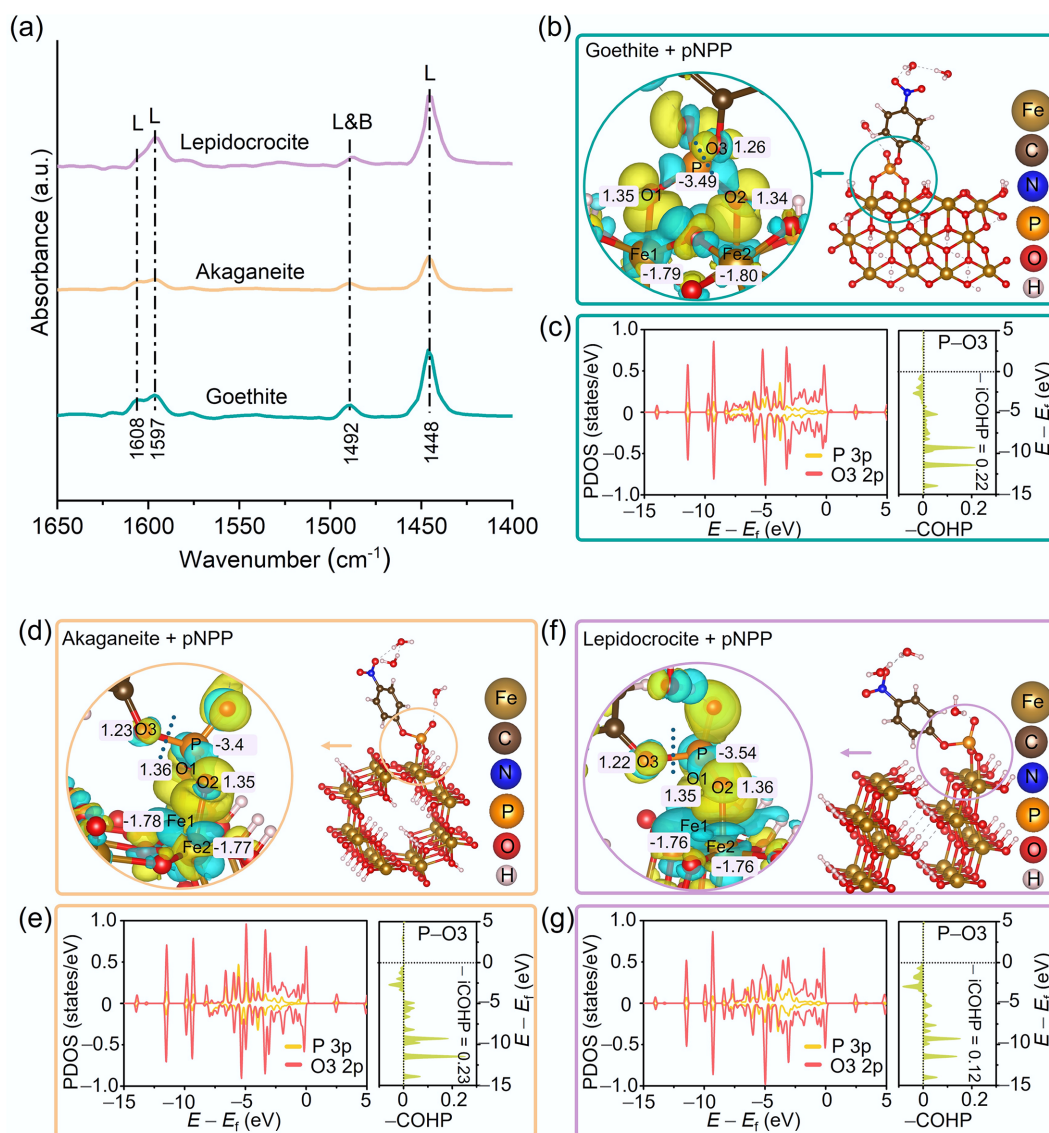


Fig. 5 (a) Py-IR spectra of goethite, akaganeite, and lepidocrocite. Charge density differences and Bader charge analyses for pNPP molecules adsorbed onto the surfaces of (b) goethite, (d) akaganeite, and (f) lepidocrocite. Yellow and cyan regions represent electron accumulation and depletion, respectively. PDOS and COHP analyses of the P 3p and O3 2p orbitals in pNPP adsorbed on (c) goethite, (e) akaganeite, and (g) lepidocrocite, respectively.

intensity; and no band corresponding to Brønsted acid sites at 1,545 cm⁻¹ was observed^[47]. Among them, lepidocrocite exhibited the highest density of Lewis acid sites (214 μmol m⁻²), significantly surpassing those of akaganeite (177 μmol m⁻²), and goethite (192 μmol m⁻²), consistent with its superior catalytic performance in pNPP hydrolysis (Figs 2 and 3, Supplementary Table S1). These acid sites play a crucial role in facilitating pNPP conversion by stabilizing electron-rich intermediates and reducing the activation energy barrier^[76].

In the catalytic system of iron oxyhydroxides, coordinatively unsaturated Fe atoms act as Lewis acid sites, engaging in electron pair sharing with the -PO₄ group of pNPP, which functions as a Lewis base, thereby driving the hydrolysis reaction^[34,48,74]. Charge density differences (Fig. 5b, d, and f) confirm this process, showing an increase in electron density on the oxygen atoms of the -PO₄ group and a corresponding partial positive charge (δ⁺) on the central P atom. This polarization effect enhances the susceptibility of the P atom to nucleophilic attack by H₂O molecules. Bader charge analysis (Fig. 5b, d, and f) further elucidates the crystalline phase-

dependent promotion of hydrolysis: the charge transfer on the P atom of pNPP adsorbed on lepidocrocite (-3.54 e) is notably greater than that observed for akaganeite (-3.49 e), and goethite (-3.44 e), indicating that lepidocrocite more effectively facilitates nucleophilic hydrolysis. Furthermore, bonding stability analysis provides deeper insight into the underlying mechanism. PDOS and COHP results (Fig. 5c, e, and g) reveal that pNPP adsorbed on lepidocrocite exhibits the weakest orbital overlap between the P 3p orbitals and the O3 2p orbitals of the ester bond. A larger proportion of covalent electrons occupies the antibonding orbitals ($E < E_f$, -COHP < 0, -iCOHP = 0.12), significantly weakening the P-O bond. This electronic configuration is identified as the intrinsic origin of the enhanced hydrolytic efficiency observed for lepidocrocite.

Conclusions

Iron oxyhydroxide nanoparticles (including goethite, akaganeite, and lepidocrocite) can effectively promote the hydrolysis of OPE pollutants

under environmentally relevant pH conditions, and their catalytic capacity is significantly affected by the crystalline phase, with surface area-normalized rate constant following the order: lepidocrocite > akaganeite > goethite. The mechanisms underlying the crystalline phase-dependent catalytic activity were elucidated by experimental and theoretical investigations. Specifically, the crystalline phase influenced both the adsorption affinity of the iron oxyhydroxide nanoparticles toward pNPP and the reactivity of the surface-bound pNPP molecules. Differences in surface charge distribution of different crystalline phases and their electronic interactions with the $-\text{PO}_4$ groups in pNPP led to variations in their adsorption affinities for pNPP (akaganeite > goethite > lepidocrocite). Moreover, crystalline phase-dependent variation in surface Lewis acidity of the iron oxyhydroxides (in the order of lepidocrocite > goethite > akaganeite) and the induced charge rearrangement of the central P atom in pNPP largely determined their overall apparent catalytic activity.

These findings deepen our understanding of the critical role of the crystalline phase in regulating the capability of metal oxide nanominerals to mediate hydrolytic transformation of OPEs, and potentially other plastic additives, which are valuable for accurately assessing the environmental behavior and potential ecological risks of these emerging pollutants. Notably, in real aquatic environments, metal oxides may interact with ions and natural organic matter, altering the surface charge, density of surface active sites, and even the crystalline phase^[61,77,78], which may dampen the crystalline phase-dependent catalytic activity observed in the simplified systems. Meanwhile, the adsorption of ions such as Zn^{2+} may promote hydrolysis reactions by metal oxides^[54]. Therefore, future research needs to elucidate how metal oxide nanominerals mediate the hydrolysis reactions under complex environmental conditions, so as to more accurately predict the environmental fate and risks of plastic additives.

Supplementary information

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Author contributions

The authors confirm their contributions to the paper as follows: investigation: Pei X, Liang Z, Chen Z; writing – original draft, methodology: Pei X, Liang Z; writing – review & editing: Pei X, Duan L, Jiang C, Alvarez PJJ, Zhang T; data curation: Pei X, Liang Z, Chen Z; funding acquisition: Pei X, Duan L, Jiang C, Zhang T; formal analysis, visualization: Pei X; visualization: Liang Z; project administration, conceptualization: Jiang C; supervision: Zhang T. All authors reviewed the results and approved the final version of the manuscript.

Data availability

The datasets used or analyzed during the current study are available from the corresponding author upon reasonable requests.

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Declarations

Competing interests

The authors declare that they have no conflict of interest.

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References

- [1] Stubbins A, Law KL, Muñoz SE, Bianchi TS, Zhu L. 2021. Plastics in the Earth system. *Science* 373:51–55
- [2] MacLeod M, Arp HPH, Tekman MB, Jahnke A. 2021. The global threat from plastic pollution. *Science* 373:61–65
- [3] Arp HPH, Kühnel D, Rummel C, MacLeod M, Potthoff A, et al. 2021. Weathering plastics as a planetary boundary threat: Exposure, fate, and hazards. *Environmental Science & Technology* 55:7246–7255
- [4] Hahladakis JN, Velis CA, Weber R, Iacovidou E, Purnell P. 2018. An overview of chemical additives present in plastics: migration, release, fate and environmental impact during their use, disposal and recycling. *Journal of Hazardous Materials* 344:179–199
- [5] Monclús L, Arp HPH, Groh KJ, Faltynkova A, Løseth ME, et al. 2025. Mapping the chemical complexity of plastics. *Nature* 643:349–355
- [6] Yan Y, Zhu F, Zhu C, Chen Z, Liu S, et al. 2021. Dibutyl phthalate release from polyvinyl chloride microplastics: Influence of plastic properties and environmental factors. *Water Research* 204:117597
- [7] Zhang QQ, Ma ZR, Cai YY, Li HR, Ying GG. 2021. Agricultural plastic pollution in China: generation of plastic debris and emission of phthalic acid esters from agricultural films. *Environmental Science & Technology* 55:12459–12470
- [8] Mali H, Shah C, Raghunandan BH, Prajapati AS, Patel DH, et al. 2023. Organophosphate pesticides an emerging environmental contaminant: pollution, toxicity, bioremediation progress, and remaining challenges. *Journal of Environmental Sciences* 127:234–250
- [9] Ma Y, Xie Z, Lohmann R, Mi W, Gao G. 2017. Organophosphate ester flame retardants and plasticizers in ocean sediments from the North Pacific to the Arctic Ocean. *Environmental Science & Technology* 51:3809–3815
- [10] Wang X, Zhu Q, Yan X, Wang Y, Liao C, et al. 2020. A review of organophosphate flame retardants and plasticizers in the environment: Analysis, occurrence and risk assessment. *Science of the Total Environment* 731:139071
- [11] Liu X, Chen D, Yu Y, Zeng X, Li L, et al. 2020. Novel organophosphate esters in airborne particulate matters: occurrences, precursors, and selected transformation products. *Environmental Science & Technology* 54:13771–13777
- [12] Wang S, Qian J, Zhang B, Chen L, Wei S, et al. 2023. Unveiling the occurrence and potential ecological risks of organophosphate esters in municipal wastewater treatment plants across China. *Environmental Science & Technology* 57:1907–1918
- [13] Rodgers TFM, Giang A, Diamond ML, Gillies E, Saini A. 2023. Emissions and fate of organophosphate esters in outdoor urban environments. *Nature Communications* 14:1175
- [14] Xie Z, Wang P, Wang X, Castro-Jiménez J, Kallenborn R, et al. 2022. Organophosphate ester pollution in the oceans. *Nature Reviews Earth & Environment* 3:309–322

- [15] Hu W, Gao P, Wang L, Hu J. 2023. Endocrine disrupting toxicity of aryl organophosphate esters and mode of action. *Critical Reviews in Environmental Science and Technology* 53:1–18
- [16] Li J, Cao H, Mu Y, Qu G, Zhang A, et al. 2020. Structure-oriented research on the antiestrogenic effect of organophosphate esters and the potential mechanism. *Environmental Science & Technology* 54:14525–4534
- [17] Yan Z, Feng C, Jin X, Wang F, Liu C, et al. 2022. Organophosphate esters cause thyroid dysfunction via multiple signaling pathways in zebrafish brain. *Environmental Science and Ecotechnology* 12:100198
- [18] Zhang Q, Yang L, Wang H, Wu C, Cao R, et al. 2024. A comprehensive evaluation of the endocrine-disrupting effects of emerging organophosphate esters. *Environment International* 193:109120
- [19] Wang Q, Lam JCW, Man YC, Lai NL S, Kwok KY, et al. 2015. Bioconcentration, metabolism and neurotoxicity of the organophorous flame retardant 1,3-dichloro 2-propyl phosphate (TDCPP) to zebrafish. *Aquatic Toxicology* 158:108–115
- [20] Louis LM, Buckley JP, Kuiper JR, Meeker JD, Hansel NN, et al. 2023. Exposures to organophosphate esters and respiratory morbidity among school-aged children with asthma. *Environmental Science & Technology* 57:6435–6443
- [21] Patisaul HB, Behl M, Birnbaum LS, Blum A, Diamond ML, et al. 2021. Beyond cholinesterase inhibition: developmental neurotoxicity of organophosphate ester flame retardants and plasticizers. *Environmental Health Perspectives* 129:105001
- [22] Zhang S, Zhao M, Li S, Yang R, Yin N, et al. 2024. Developmental toxicity assessment of neonicotinoids and organophosphate esters with a human embryonic stem cell- and metabolism-based fast-screening model. *Journal of Environmental Sciences* 137:370–381
- [23] Witchev SK, Sutherland V, Collins B, Roberts G, Shockley KR, et al. 2023. Reproductive and developmental toxicity following exposure to organophosphate ester flame retardants and plasticizers, triphenyl phosphate and isopropylated phenyl phosphate, in Sprague Dawley rats. *Toxicological Sciences* 191:374–386
- [24] Zhang Y, Qin H, Li B, Yu Z, Zu B, et al. 2025. A Novel organophosphate ester, tris (2,4-ditert-butylphenyl) phosphate, induced reproductive toxicity in male zebrafish at environmentally relevant concentrations. *Environmental Science & Technology* 59:279–290
- [25] Li J, He J, Li Y, Liu Y, Li W, et al. 2019. Assessing the threats of organophosphate esters (flame retardants and plasticizers) to drinking water safety based on USEPA oral reference dose (RfD) and oral cancer slope factor (SFO). *Water Research* 154:84–93
- [26] Zhang X, Lu Z, Ren X, Chen X, Zhou X, et al. 2021. Genetic comprehension of organophosphate flame retardants, an emerging threat to prostate cancer. *Ecotoxicology and Environmental Safety* 223:112589
- [27] Liu M, Li A, Meng L, Zhang G, Guan X, et al. 2022. Exposure to novel brominated flame retardants and organophosphate esters and associations with thyroid cancer risk: a case-control study in eastern China. *Environmental Science & Technology* 56:17825–17835
- [28] Sun B, Zhou C, Zhu M, Wang S, Zhang L, et al. 2024. Leaching kinetics and bioaccumulation potential of additive-derived organophosphate esters in microplastics. *Environmental Pollution* 347:123671
- [29] Castro-Jiménez J, Aminot Y, Bely N, Pollono C, Idjaton BIT, et al. 2024. Organophosphate ester additives and microplastics in benthic compartments from the Loire estuary (French Atlantic coast). *Marine Pollution Bulletin* 201:116256
- [30] Chen Y, Chen Q, Zhang Q, Zuo C, Shi H. 2022. An overview of chemical additives on (micro) plastic fibers: occurrence, release, and health risks. *Reviews of Environmental Contamination and Toxicology* 260:22
- [31] Onoja S, Nel HA, Abdallah MAE, Harrad S. 2022. Microplastics in freshwater sediments: Analytical methods, temporal trends, and risk of associated organophosphate esters as exemplar plastics additives. *Environmental Research* 203:111830
- [32] Chen W, Gong Y, McKie M, Almuhtaram H, Sun J, et al. 2022. Defining the chemical additives driving in vitro toxicities of plastics. *Environmental Science & Technology* 56:14627–14639
- [33] Fu Z, Xie HB, Elm J, Liu Y, Fu Z, et al. 2022. Atmospheric autoxidation of organophosphate esters. *Environmental Science & Technology* 56:6944–6955
- [34] Fang Y, Kim E, Strathmann TJ. 2018. Mineral- and base-catalyzed hydrolysis of organophosphate flame retardants: potential major fate-controlling sink in soil and aquatic environments. *Environmental Science & Technology* 52:1997–2006
- [35] Zhou X, Liang Y, Ren G, Zheng K, Wu Y, et al. 2020. Biotransformation of tris (2-chloroethyl) phosphate (TCEP) in sediment microcosms and the adaptation of microbial communities to TCEP. *Environmental Science & Technology* 54:5489–5497
- [36] Hochella MF Jr, Mogk DW, Ranville J, Allen IC, Luther GW, et al. 2019. Natural, incidental, and engineered nanomaterials and their impacts on the earth system. *Science* 363:aa8299
- [37] Huang X, Hou X, Zhang X, Rosso KM, Zhang L. 2018. Facet-dependent contaminant removal properties of hematite nanocrystals and their environmental implications. *Environmental Science-Nano* 5:1790–1806
- [38] Olsson R, Giesler R, Loring JS, Persson P. 2012. Enzymatic hydrolysis of organic phosphates adsorbed on mineral surfaces. *Environmental Science & Technology* 46:285–291
- [39] Gorski CA, Nurmi JT, Tratnyek PG, Hofstetter TB, Scherer MM. 2010. Redox behavior of magnetite: Implications for contaminant reduction. *Environmental Science & Technology* 44:55–60
- [40] Liu C, Li F, Chen M, Liao C, Tong H, et al. 2017. Adsorption and stabilization of lead during Fe(II)-catalyzed phase transformation of ferrihydrite. *Acta Chimica Sinica* 75:621–628
- [41] Aeppli M, Voegelin A, Gorski CA, Hofstetter TB, Sander M. 2018. Mediated electrochemical reduction of iron (oxyhydr)- oxides under defined thermodynamic boundary conditions. *Environmental Science & Technology* 52:560–570
- [42] Huang X, Hou X, Song F, Zhao J, Zhang L. 2016. Facet-dependent Cr(VI) adsorption of hematite nanocrystals. *Environmental Science & Technology* 50:1964–1972
- [43] Yan W, Jing C. 2018. Molecular insights into glyphosate adsorption to goethite gained from ATR-FTIR, two-dimensional correlation spectroscopy, and DFT study. *Environmental Science & Technology* 52:1946–1953
- [44] Francisco PCM, Sato T, Otake T, Kasama T, Suzuki S, et al. 2018. Mechanisms of Se(IV) Co-precipitation with ferrihydrite at acidic and alkaline conditions and its behavior during aging. *Environmental Science & Technology* 52:4817–4826
- [45] Ardo SG, Nélieu S, Ona-Nguema G, Delarue G, Brest J, et al. 2015. Oxidative degradation of nalidixic acid by nano-magnetite via Fe²⁺/O₂⁻ mediated reactions. *Environmental Science & Technology* 49:4506–4514
- [46] Zhao P, Begg JD, Zavarin M, Tumey SJ, Williams R, et al. 2016. Plutonium(IV) and (V) sorption to goethite at sub-femtomolar to micromolar concentrations: redox transformations and surface precipitation. *Environmental Science & Technology* 50:6948–6956
- [47] Lin J, Hu S, Liu T, Li F, Peng L, et al. 2019. Coupled kinetics model for microbially mediated arsenic reduction and adsorption/desorption on iron oxides: role of arsenic desorption induced by microbes. *Environmental Science & Technology* 53:8892–8902
- [48] Li T, Zhong W, Jing C, Li X, Zhang T, et al. 2020. Enhanced hydrolysis of p-nitrophenyl phosphate by iron (hydr) oxide nanoparticles: Roles of exposed facets. *Environmental Science & Technology* 54:8658–8667
- [49] Dannenberg A, Pehkonen SO. 1998. Investigation of the heterogeneously catalyzed hydrolysis of organophosphorus pesticides. *Journal of Agricultural and Food Chemistry* 46:325–334
- [50] Wu D, Huang S, Zhang X, Ren H, Jin X, et al. 2021. Iron minerals mediated interfacial hydrolysis of chloramphenicol antibiotic under limited moisture conditions. *Environmental Science & Technology* 55:9569–9578
- [51] Jin X, Wu D, Liu C, Huang S, Zhou Z, et al. 2022. Facet effect of hematite on the hydrolysis of phthalate esters under ambient humidity conditions. *Nature Communications* 13:6125
- [52] Pei X, Jiang C, Chen W. 2019. Enhanced hydrolysis of 1,1,2,2-tetrachloroethane by multi-walled carbon nanotube/TiO₂ nanocomposites: the synergistic effect. *Environmental Pollution* 255:113211
- [53] Jin X, Wu D, Ling J, Wang C, Liu C, et al. 2019. Hydrolysis of chloramphenicol catalyzed by clay minerals under nonaqueous conditions. *Environmental Science & Technology* 53:10645–10653

- [54] Sheng F, Ling J, Wang C, Jin X, Gu X, et al. 2019. Rapid hydrolysis of penicillin antibiotics mediated by adsorbed zinc on goethite surfaces. *Environmental Science & Technology* 53:10705–10713
- [55] Lei Y, Mulchandani P, Wang J, Chen W, Mulchandani A. 2005. Highly sensitive and selective amperometric microbial biosensor for direct determination of p-nitrophenyl-substituted organophosphate nerve agents. *Environmental Science & Technology* 39:8853–8857
- [56] Mahaninia MH, Wilson LD. 2016. Modular cross-linked chitosan beads with calcium doping for enhanced adsorptive uptake of organophosphate anions. *Industrial & Engineering Chemistry Research* 55:11706–11715
- [57] Akçay M. 2005. The surface acidity and characterization of Fe-montmorillonite probed by in situ FT-IR spectroscopy of adsorbed pyridine. *Applied Catalysis A: General* 294:156–160
- [58] Pei X, Wang W, Chen Z, Liu K, Liang Z, et al. 2024. Metal heteroatoms significantly enhance efficacy of TiO₂ nanomaterials in promoting hydrolysis of organophosphates: Implications for mitigating pollution of plastic additives. *The Science of the Total Environment* 957:177548
- [59] Kumar KV, Porkodi K, Rocha F. 2008. Langmuir–Hinshelwood kinetics – A theoretical study. *Catalysis Communications* 9:82–84
- [60] Sakamoto Y, Noda Y, Ohno K, Koike K, Fujii K, et al. 2019. First principles calculations of surface dependent electronic structures: a study on β -FeOOH and γ -FeOOH. *Physical Chemistry Chemical Physics* 21:18486–18494
- [61] Lv J, Zhang S, Wang S, Luo L, Cao D, et al. 2016. Molecular-scale investigation with ESI-FT-ICR-MS on fractionation of dissolved organic matter induced by adsorption on iron oxyhydroxides. *Environmental Science & Technology* 50:2328–2336
- [62] Wu S, Lu J, Ding Z, Li N, Fu F, et al. 2016. Cr(VI) removal by mesoporous FeOOH polymorphs: performance and mechanism. *RSC Advances* 6:82118–82130
- [63] Ding J, Shen L, Yan R, Lu S, Zhang Y, et al. 2020. Heterogeneously activation of H₂O₂ and persulfate with goethite for bisphenol A degradation: A mechanistic study. *Chemosphere* 261:127715
- [64] Thommes M, Kaneko K, Neimark AV, Olivier JP, Rodriguez-Reinoso F, et al. 2015. Physisorption of gases, with special reference to the evaluation of surface area and pore size distribution (IUPAC Technical Report). *Pure and Applied Chemistry* 87:1051–1069
- [65] Wang H, Cao S, Kang F, Chen R, Liu H, et al. 2014. Effects of Al substitution on the microstructure and adsorption performance of α -FeOOH. *Journal of Alloys and Compounds* 606:117–123
- [66] Qin M, Lu B, Feng S, Zhen Z, Chen R, et al. 2019. Role of exposed facets and surface OH groups in the Fenton-like reactivity of lepidocrocite catalyst. *Chemosphere* 230:286–293
- [67] Liang Z, Liu K, Li Y, Liu Y, Jiang C, et al. 2025. Oxygen vacancies boost the efficacy of MnO₂ nanoparticles in catalyzing the hydrolytic degradation of organophosphate esters: implications for managing plastic additive pollution. *Environmental Science-Nano* 12:1364–1374
- [68] Li Y, Liu K, Guo Y, Wang H, Liang Z, et al. 2025. Mechanistic insights into facet-dependent catalytic hydrolysis of organophosphate ester by α -MnO₂ nanorods. *Separation and Purification Technology* 360:130940
- [69] Luengo C, Brigante M, Antelo J, Avena M. 2006. Kinetics of phosphate adsorption on goethite: Comparing batch adsorption and ATR-IR measurements. *Journal of Colloid and Interface Science* 300:511–518
- [70] Luo L, Lv J, Chen Z. 2018. Synchrotron infrared microspectroscopy reveals the roles of aliphatic and aromatic moieties in sorption of nitroaromatic compounds to soils. *Science of the Total Environment* 624:210–214
- [71] Barja BC, Tejedor-Tejedor MI, Anderson MA. 1999. Complexation of methylphosphonic acid with the surface of goethite particles in aqueous solution. *Langmuir* 15:2316–2321
- [72] Parikh SJ, Chorover J. 2006. ATR-FTIR spectroscopy reveals bond formation during bacterial adhesion to iron oxide. *Langmuir* 22:8492–8500
- [73] Jiang X, Zhou B, Yang W, Chen J, Miao C, et al. 2024. Precise coordination of high-loading Fe single atoms with sulfur boosts selective generation of nonradicals. *Proceedings of the National Academy of Sciences of the United States of America* 121:e2309102121
- [74] Li T, Ju Y, Du T, Jiang C, Zhang T, et al. 2024. Anatase TiO₂ nanomaterials are much more effective in enhancing hydrolysis of organophosphorus compounds than their rutile counterparts. *Environmental Science-Nano* 11:2447–2456
- [75] He L, Jing L, Luan Y, Wang L, Fu H. 2014. Enhanced visible activities of α -Fe₂O₃ by coupling N-doped graphene and mechanism insight. *ACS Catalysis* 4:990–998
- [76] Bhattacharyya K, Danon A, Vijayan BK, Gray KA, Stair PC, et al. 2013. Role of the surface Lewis acid and base sites in the adsorption of CO₂ on titania nanotubes and platinized titania nanotubes: An in situ FT-IR study. *Journal of Physical Chemistry C* 117:12661–12678
- [77] Otero-Fariña A, Fiol S, Arce F, Antelo J. 2017. Effects of natural organic matter on the binding of arsenate and copper onto goethite. *Chemical Geology* 459:119–128
- [78] Schulz K, Wisawapipat W, Barmettler K, Grigg ARC, Kubeneck LJ, et al. 2024. Iron oxyhydroxide transformation in a flooded rice paddy field and the effect of adsorbed phosphate. *Environmental Science & Technology* 58:10601–10610



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