

Impact of fertilization and tillage practices on transformations of carbon, essential plant nutrients and microbial biota composition in soils: a review

Salwinder Singh Dhaliwal^{1*}, Arvind Kumar Shukla², Sanjib Kumar Behera³, Sarwan Kumar Dubey⁴, Sandeep Sharma¹, Mehakpreet Kaur Randhawa¹, Gagandeep Kaur⁵, Sohan Singh walia⁶ and Amardeep Singh Toor¹

¹ Department of Soil Science, Punjab Agricultural University, Ludhiana 141004, India

² Rajmata Vijayaraje Scindia Krishi Vishwa Vidyalaya, Gwalior 474002, India

³ Indian Institute of Soil Science (IISS), Bhopal 462038, India

⁴ ICAR-Indian Institute of Soil and Water Conservation, CSSRI, Karnal 132001, India

⁵ Department of Crop and Soil Sciences, Washington State University, Pullman, WA 99164, USA

⁶ School of Organic Farming, Punjab Agricultural University, Ludhiana, Punjab 141004, India

* Corresponding author, E-mail: ssdhalwal@pau.edu

Abstract

Soil management approaches have been advocated to modify the soil fertility parameters for higher agricultural production through different land systems. The present review examines the influence of organic/inorganic fertilizers and tillage practices through transformations in regulating the nutrient status, microbial components, and soil organic carbon. Fertilization along with different tillage practices have been found to affect the available plant nutrient content including macronutrients, secondary nutrients, and micronutrients. The review investigation also showed that, compared to inorganic fertilizers (INF), application of compost enhanced plant available macronutrients (N, P, K), micronutrients (Zn, Cu, Fe, Mn) and soil organic carbon (SOC) with different tillage practices. Through different land systems, transformation of the plant available macronutrients, micronutrients and microbial compositions showed their enhancement. Microbial parameters viz. microbial biodiversity, microbial biomass carbon (MBC), potentially mineralizable nitrogen (PMN), microbial biomass nitrogen (MBN) and microbial respiration reported increase. Soil organic carbon and aggregate distribution in the soil and the aggregate-associated organic carbon and physical fractions of SOC have also been reviewed. Among different tillage systems, the reduced tillage with residue incorporation and no-tillage (zero tillage) with residue mulching, significantly enhanced carbon sequestration in soil aggregates in comparison to conventional tillage with residue removal treatments. The practice of zero tillage improved dissolved organic carbon and MBC in light and heavy fractions of carbon in the upper layers of soil.

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Introduction

Agricultural management practices impact soil physico-chemical properties to a remarkable extent. Degradation of soil health has led to a contraction in agricultural production and soil biodiversity particularly due to conventional farming practices, indiscriminate use of inorganic fertilizers (INF) and inadequate input of residues^[1]. Organic or inorganic fertilizers have been regarded as a critical component of agriculture to accomplish global food security goals^[2]. The exogenous supply of fertilizers could easily alter soil properties by restoring the nutrients that have been absorbed by the plants^[2]. Thus, implementing adequate nutrient management strategies could boost plant yield and sustain plant health. Tillage affects the soil, especially for crop production and consequently affects the agro-ecosystem functions. This involves the mechanical manipulation of the soil to modify soil attributes like soil water retention, evapotranspiration, and infiltration processes for better crop production. Thus, tillage practices coupled with fertilizer inputs may prove a viable strategy to improve soil

health components such as nutrient status, biodiversity, and organic carbon.

Soil serves as a major reservoir of nutrients for sustainable crop production. Intensive cultivation due to growing population burden has led to the decline of soil nutrient status that has adversely affected agricultural production. Various researchers have assessed the soil nutrient budget and the reasons behind decline of nutrient content in soil^[3]. Soil management strategies have assisted in overcoming this problem to a greater extent. Tillage practices redistribute soil fertility and improve plant available nutrient content due to soil perturbations^[4]. Different tillage and fertilization practices alter soil nutrient cycling over time^[5]. Fertilization is an important agricultural practice which is known to increase nutrient availability in soil as well as plants^[6]. A report has been compiled by Srivastava et al.^[7], which assessed the effectiveness of different fertilizers on soil nutrient status in Indian soils.

Soil biota has a vital role in the self-regulating functions of the soil to maintain soil quality which might reduce the reliance on anthropogenic activities. Soil microbial activities are

sensitive to slight modifications in soil properties and could be used as an index of soil health^[8]. Maintenance of microbial activity is essential for soil resilience as they influence a diverse range of parameters and activities including soil structure formation, soil SOM degradation, bio-geochemical cycling of nutrients etc.^[9]. Various researchers have identified microbial parameters like microbial biomass carbon (MBC), potentially mineralizable nitrogen (PMN), soil respiration, microbial biomass nitrogen (MBN), and earthworm population as potential predictors of soil quality. Geisseler & Scow^[10] have compiled a review on the affirmative influence of long-term mineral fertilization on the soil microbial community.

Being the largest terrestrial carbon (C) reservoir, soil organic carbon (SOC) plays a significant role in agricultural productivity, soil quality, and climate change mitigation^[11]. Manure addition, either solely or along with INF augments SOC content which helps in the maintenance and restoration of SOM more effectively as compared to the addition of INF alone^[12]. Enhancement of recalcitrant and labile pools of SOC could be obtained through long-term manure application accentuating the necessity of continuous organic amendments for building up C and maintaining its stability^[13]. Generally, compared with manure addition, INF application is relatively less capable of raising SOC and its labile fractions^[14]. Alteration in SOC content because of management strategies and/or degradation or restoring processes is more prominent in the labile fraction of soil C^[15]. Several fractions of soil C play vital roles in food web and nutrient cycles in soils besides influencing many biological properties of soil^[16]. Thus, monitoring the response of SOC and its fractions to various management practices is of utmost importance.

A positive impact on SOC under manure application coupled with INF in rice-wheat systems has been reported, as compared to sole applications of INFs^[17]. Although ploughing and other mechanical disturbances in intensive farming cause rapid OM breakdown and SOC loss^[18], additional carbon input into the soil through manure addition and rational fertilization increases carbon content^[13]. Wei et al.^[19] in light sandy loam soil of China found that the inclusion of crop straw together with inorganic N, P and K fertilizers showed better results for improving soil fertility over sole use of inorganic fertilizers. Zhu et al.^[20] studied the influence of soil C through wheat straw, farmyard manure (FYM), green manure, and rice straw on plant growth, yield, and various soil properties and found that the recycling of SOM under intensive cultivation is completely reliant on net OM input and biomass inclusion. However, most of the studies on residue management and organically managed systems could not provide clear views regarding the relations between the quality of OM inputs and biological responses towards it. The disintegration of soil aggregates due to ploughing, use of heavy machinery, and residue removal has been reported widely under conventional tillage (CT) practices^[21]. On the contrary, improvement in SOC stabilization has also been observed by some scientists^[22]. Under CT, the disintegration of macro-aggregates into micro-aggregates is a prominent phenomenon, while conservation tillage has been identified as a useful practice for increasing macro-aggregates as well as carbon sequestration in agricultural soils^[23]. By and large, the ploughing depth (0–20 cm) is taken into consideration for evaluating the impact of tillage and straw retention on soil aggregation^[24], while degradation in deeper layers of soil is

becoming a major constraint towards soil quality together with crop yield^[25].

Hence, the present review would be useful in determining how tillage practices and inorganic and organic fertilization impact nutrient availability in the soil, microbial composition and SOC fractions besides stocks under different land uses.

Effect of fertilization and tillage practices on the availability of essential plant nutrients

Agricultural production is greatly influenced by nutrient availability and thus nutrient management is required for sustaining higher yields of crop. The term 'nutrient availability' refers to the quantity of nutrients in chemical forms accessible to plant roots or compounds likely to be converted to such forms throughout the growing season *in lieu* of the total amount of nutrients in the soil. For optimum growth, different crops require specifically designed nutrient ratios. Plants need macronutrients [nitrogen (N), phosphorus (P), potassium (K) in higher concentrations], secondary nutrients [calcium (Ca), magnesium (Mg), sulphur (S) in moderate amounts as compared to macronutrients] and Micronutrients [Zn (zinc), Fe (iron), Cu (copper), B (boron), Mn (manganese), Mo (molybdenum) in smaller amounts] for sustainable growth and production^[26]. Fertilizers assist the monitoring of soil nutrient levels by direct addition of required nutrients into the soil through different sources and tillage practices may alter the concentration of available plant nutrients through soil perturbations. Various studies on the influence of fertilization and tillage practices on available plant nutrients have been discussed below.

Plant available macronutrients (N, P, K) in soil

Yue et al.^[27] reported that long-term fertilization through manure/INF improved the macronutrient content of *Ultisol* soil in China. Two doses of NPK (2NPK) considerably improved soil properties over a single dose (NPK). Combined application (NPK + OM) resulted in higher hydrolysable N and available P over the sole OM application. The total K content was higher under the treatments NPK, 2NPK and NPK + OM than sole OM treatment, whereas available K was higher in treatments NPK + OM and 2NPK over the sole OM and NPK. Likewise, OM, INF, and OM + INF were evaluated for their potential to regulate the soil macronutrient dynamics. Organic manure significantly improved the soil N content, whereas INF showed comparable results to that of the control treatment. Besides, all the treatments improved available P and exchangeable K concentration^[28].

Hasnain et al.^[29] performed comparative studies of different ratios of INF + compost and different application times for the chemical N fertilizer on silty loamy soils of China. The available nitrogen and phosphorous content were greater in conjoint OM + INF application over the bare INF and control application irrespective of N application time. Soil quality substantially improved with increasing ratio of compost and 70:30 (INF to compost ratio) was found to be most suitable to maintain soil fertility and nutrient status. Another study by Liu et al.^[30] reported the superior effects of NPK + pig manure and NPK + straw to improve soil available P and K over the control and sole NPK treatments. However, total N concentration did not exhibit any significant variation under any treatment.

Shang et al.^[31] accounted the positive impact of vermicompost and mushroom residue application on grassland soil fertility in China. The addition of organic manures improved

available P and K content to a considerable extent. Under moisture-deficit winter wheat-fallow rotation, another study quantified the influence of residue management approaches and fertilizer rate on nutrient accrual. Residue burning resulted in no decline in soil macronutrient content, whereas the perpetual addition of FYM for 84 years significantly improved total N and extractable K and P concentration. Thus, residue incorporation along with FYM application may prove beneficial in reducing the temporal macronutrient decline^[32].

Ge et al.^[33] examined the effects of NPK and NPK along with manure (NPKM) addition on the macronutrient status of *Hapli-Udic Cambisol* soil. The NPKM application resulted in the highest increase in total N, available-P and K concentration as compared to NPK and control. Likewise, mineral fertilization reduction and partial substitution with organic amendments have posed a significant influence on soil macronutrient status. Soil available P and K decreased after INF reduction^[34]. Chen et al.^[35] evinced that integrated application of manure and mineral fertilizers to red clay soil (typical *Ultisols*) improved hydrolyzed nitrogen and available P due to an increase in the decomposition of organic matter (OM) and N bio-fixation than sole mineral fertilizers and control.

A long-term experiment was carried out by Shiwakoti et al.^[36] to ascertain the influence of N fertilization and tillage on macronutrient dynamics in soil. Nitrogen fertilization produced higher crop biomass which might have improved total N and P concentration in soil. Moreover, the reduced interaction between soil colloids and residue or greater cation exchange sites due to tillage practices could have augmented K concentration in 0–10 cm soil depth. Likewise, among tillage systems combined organic (poultry manure) and inorganic (lime and fertilizers) fertilization, no-tillage, and reduced tillage with organic fertilization resulted in higher availability of P owing to minimal disturbance of soil which decreases contact surface between phosphate ions and adsorption sites. Greater losses of K in runoff water under NT resulted in lower K availability under NT than CT^[37].

The influence of tillage systems on soil nutrient dynamics showed that minimal soil disturbances under zero tillage prohibited redistribution of soil nutrients and resulted in the highest available N, P, and K in the surface soil^[38]. The influence of tillage timing on soil macronutrient status has also been assessed under tillage treatments that are fall tillage (FT), spring tillage (ST), no tillage (NT), and disk/chisel tillage (DT/CT) on mixed mesic Typic *Haploxerolls* soil. All the tillage systems differed in the quantity of residues generated. Thus, variation in the decomposition of crop residue and mineralization of SOM resulted in variable rates of nutrient release. The FT and ST had the highest N content over DT/CT and NT systems at corresponding depth. The N content also decreased with soil depth irrespective of tillage treatment. The available P and extractable K were highest under NT at the top 10 cm soil depth and increased over time^[39]. Residue management in combination with tillage treatments (ST and CT) has been reported to affect the soil macronutrient status in Bangladesh. Tillage treatments enhanced the total N content to a considerable extent. Moreover, 3 years of residue retention led to a higher concentration of total N, available P and K in the soil.

Plant available micronutrients (Zn, Cu, Fe, Mn) in soil

The combinations of N, P, and K in different ratios together with two rates of organic fertilizer (OF) applied on the *aquic*

Inceptisol having sandy loam texture influenced the micronutrient status of the soil^[40]. Soil Zn content decreased with time when no fertilizer was applied as compared to organic fertilizer (OF) application. The mineral fertilizer treatments led to a substantial increase in DTPA-extractable micronutrients in the soil. The higher micronutrient concentration due to higher OM highlights the importance of maintaining OM for soil fertility and higher crop production. Further studies revealed that long-term application of sole N fertilizers led to a significant decline in total Zn and Cu, whereas Mn and Fe status improved through atmospheric deposition. Phosphorus and OF addition along with straw incorporation markedly increased total Zn, Cu, Fe, and Mn. The DTPA-extractable Mn, Zn, Fe, and Cu were also higher in OF treatment, thus demonstrating the beneficial effects of constant OM application for maintaining the nutrient status of soil^[41].

López-Fando & Pardo^[42] quantified the impact of various tillage practices including NT, CT, minimum tillage (MT), and zone-tillage (ZT) on soil micronutrient stocks. Tillage systems did exhibit a significant influence on plant available Fe stocks in the topsoil; however, diminished with depth under ZT, NT and MT. Manganese was higher in NT and ZT at all depths and increased with soil depth. Zinc was highest under NT and other results did not vary significantly as in the case of Cu. The SOC levels were also found to be responsible to affect micronutrients due to tillage practices. Likewise, in *Calcitridic Haploxeralf* soil the distribution of soil micronutrients (Zn, Mn, Fe, Cu) was ascertained under different tillage practices (CT, MT, and NT). The micronutrient status was highest under NT in the upper layers due to the higher SOC level^[43].

Sharma & Dhaliwal^[44] determined that the combined application of nitrogen and rice residues facilitated the transformation of micronutrients (Zn, Mn, Fe, Cu). Among different fractions, the predominant fractions were crystalline Fe bound in Zn, Mn, and Cu and amorphous Fe oxide in Fe with 120 kg N ha⁻¹ and 7.5-ton rice residue incorporation. The higher content of occluded fractions adduced the increment in cationic micronutrient availability in soil with residue incorporation together with N fertilization due to increased biomass. Rice straw compost along with sewage sludge (SS) and INF also affected the micronutrient availability under the RW cropping system. Nitrogen fertilization through inorganic fertilizers and rice straw compost and sewage sludge (50% + 50%) improved soil micronutrient status due to an increase in SOM over sole NPK fertilizers^[45]. Earlier, Dhaliwal et al.^[46] in a long-term experiment determined that different combinations of NPK along with biogas slurry as an organic source modified the extractable micronutrient status of the soil.

A comparative study was carried out by Dhaliwal et al.^[47] to ascertain the long-term impact of agro-forestry and rice-wheat systems on the distribution of soil micronutrients. The DTPA-extractable and total Cu, Zn, Fe, and Mn were greater in the RW system due to the reduced conditions because of rice cultivation. Under the RW system Zn removal was higher which was balanced by continuous Zn application. The higher availability of Fe under the RW system was due to reduced conditions. Contrarily, Mn was greater under the agro-forestry system owing to nutrient recycling from leaf litter.

The long-term impact of integrated application of FYM, GM, WCS (wheat-cut straw) and INF on the soil micronutrients (Zn, Mn, Cu, and Fe) have been studied by Dhaliwal et al.^[48]. The FYM application substantially improved DTPA-extractable Zn

status followed by GM and WCS, whereas Cu content was maximum in the plots with OM application. The highest Fe concentration was recorded in treatment in which 50% recommended N supplied through FYM. This could be ascribed to the release of micronutrients from OM at low soil pH.

Shiwakoti et al.^[49] studied the dual effects of tillage methods (MP, DP, SW) and variable rates of N (0, 45, 90, 135 and 180 kg ha⁻¹) on the distribution of micronutrients under a moisture-deficit winter wheat-fallow system. The soil Mn content was highest under the DP regime. Inorganic N application reduced Cu content in the soil. Comparative studies with adjacent undisturbed grass pasture indicated the loss of Zn and Cu to a significant extent. Thus, DP along with nitrogen added through inorganic fertilizers could improve micronutrient concentration in the soil. Moreover, the results implied that long-term cultivation with nitrogen fertilization and tillage results in the decline of essential plant nutrients in the soil. Thus, organic amendments along with INF may prove an effective approach to increase soil micronutrient content. In another study conducted by Lozano-García & Parras-Alcántara^[50] tillage practices such as NT under apple orchard, CT with the wheat-soybean system and puddling (PD) in the rice-rice cropping system were found to affect nutrient status. Under CT, Cu content was lowest and Zn content was highest. On the contrary, puddling caused an increase in Fe and Mn concentration owing to the dispersion of soil aggregates which reduced the percolation of water and created an anaerobic environment thereby enhancing the availability of Fe and Mn.

Plant available secondary nutrients (Ca, Mg, S) in soil

Tillage practices along with gypsum fertilization have been known to affect secondary nutrient concentrations in soil. In a long-term experiment, FYM application showed maximum response to increased S concentration due to the maximum addition of OM through FYM over other treatments as S is an essential component of OM and FYM^[32]. Higher Mg content was recorded in FYM and pea vine treatments because the application of organic matter through organic manure or pea vines outright led to Mg accrual. The lower Mg concentration in topsoil than the lower layers was due to the competition between Mg and K for adsorbing sites and thus displacement of Mg by K. Han et al.^[28] while ascertaining the impact of organic manures and mineral fertilizers (NPK) on soil chemical attributes determined that INF application reduced exchangeable calcium, whereas no significant changes were exhibited in the magnesium concentrations. The OM application significantly increased both the calcium and magnesium concentrations in the soil.

While ascertaining the effect of different tillage treatments such as CT, NT, and MT on exchangeable and water-soluble cations, Lozano-García & Parras-Alcántara^[50] recorded that NT had greater content of exchangeable Ca²⁺ and Mg²⁺ than MT and CT. The exchangeable Ca²⁺ decreased with depth, however, opposite results were observed for Mg²⁺ which might be due to the higher uptake of Mg²⁺ by the crop. On another note, there might be the existence of Mg²⁺-deficient minerals on the surface horizon. Alam et al.^[51] studied the temporal effect of tillage systems on S distribution in the soil and observed that available S was 19%, 31%, and 34% higher in zero tillage than in minimum tillage, conventional tillage, and deep tillage, respectively.

Kumar et al.^[38] appraised the impact of tillage systems on surface soil nutrient dynamics under the following conditions: conventional tillage, zero till seeding with bullock drawn, conventional tillage with bullock drawn seeding, utera cropping and conservation tillage seeding with country plough and observed that tillage had a significant impact on the available S content. Compared with conventional tillage, zero and minimum tillage had higher S content as there was none or limited tillage operations which led to the accumulation of root stubble in the soil that decomposed over time and increased S concentration.

Effect of fertilization and tillage practices on the composition of the soil microbial biota

Soil is considered a hotspot for microbial biodiversity which plays an important role in building a complex link between plants and soil. The microbial components exhibit dynamic nature and, therefore, are characterized as good indicators of soil quality^[52]. These components include MBC, MBN, PMN and microbial respiration which not only assist in biological transformations like OM conversion, and biological nitrogen fixation but also increase nutrient availability for crop uptake. Management strategies such as fertilizer inputs and tillage practices may exert beneficial effects on soil biota as discussed below.

Microbial community and biodiversity in the soil

Soil is an abode to a considerable portion of global biodiversity. This biodiversity not only plays a pivotal role in regulating soil functions but also provides a fertile ground for advancing global sustainability, especially agricultural ventures. Thus, the maintenance of soil biodiversity is of paramount importance for sustaining ecosystem services. Soil biodiversity is the diverse community of living creatures in the soil that interact not only with one another but also with plants and small animals to regulate various biological activities^[53]. Additionally, it increases the fertility of soil by converting organic litter to SOM thereby enhancing SOC content. Thus, the SOM measures the number and activity of soil biota. Furthermore, the quality and amount of SOC, as well as plant diversity have a considerable impact on the soil microbial community structure^[54].

Dangi et al.^[55] ascertained the impact of integrated nutrient management and biochar on soil microbial characteristics and observed that soil amended with biochar or the addition of organic manures influenced microbial community composition and biomass and crop yield. After two years, the higher rates of biochar significantly enhanced the levels of gram-positive and gram-negative bacterial phospholipid fatty acid (PLFA), total arbuscular mycorrhizal fungal (AMF) than lower rates, unfertilized and non-amended soil. Luan et al.^[56] conducted a comparison study in a greenhouse to assess the effects of various rates of N fertilizer and kinds (inorganic and organic) on enzyme activities and soil microbial characteristics. Microbial growth (greater total PLFAs and microbial biomass carbon) and activity were promoted by manure substitution of mineral fertilizer, particularly at a higher replacement rate. On account of lower response in bacterial over fungal growth, manure addition led to a greater fungi/bacteria ratio. Furthermore, manure application significantly enhanced microbial communities, bacterial stress indicators and functional diversity. Lazcano et al.^[57] determined the influence of different fertilization strategies on microbial community structure and

function, soil biochemical properties and crop yield three months after addition of fertilizer. The integrated fertilizer regimes augmented microbial growth with improved enzyme activity as compared to sole inorganic amendments. Bacterial growth showed variable response with variation in fertilizer regime used whereas fungal growth varied with the amount of fertilizer added. Compared to mineral fertilizers, manure application led to a rapid increase in PLFA biomarkers for gram-negative bacteria. The organic amendments exhibited significant effects even at small concentration of the total quantity of nutrients applied through them; thus, confirming the viability of integrated fertilizer strategies in the short term.

Kamaa et al.^[58] assessed the long-term effect of crop manure and INF on the composition of microbial communities. The organic treatments comprised of maize (*Zea mays*) stover (MS) at 10 t ha⁻¹ and FYM @ 10 t ha⁻¹, INF treatments (120 kg N, 52.8 kg P-N₂P₂), integrated treatments (N₂P₂ + MS, N₂P₂ + FYM), fallow plot and control. The treatment N₂P₂ exhibited unfavourable effects on bacterial community structure and diversity that were more closely connected to the bacterial structure in control soils than integrated treatments or sole INF. In N₂P₂, fungal diversity varied differently than bacterial diversity but fungal diversity was similar in the N₂P₂ + FYM and N₂P₂ + MS-treated plots. Thus, the total diversity of fungal and bacterial communities was linked to agroecosystem management approaches which could explain some of the yield variations observed between the treatments. Furthermore, a long-term experiment was performed by Liu et al.^[59] to study the efficiency of pig manure and compost as a source for N fertilization and found unique prokaryotic communities with variable abundance of *Proteobacteria* under compost and pig manure treatments.

Recently, Li et al.^[60] assessed the influence of different tillage practices (no-tillage, shallow tillage, deep tillage, no-tillage with straw retention, shallow tillage with straw retention and deep tillage with straw retention) on microbial communities and observed that tillage practices improved the bacterial Shannon index to a greater extent over the no-tillage plots in which the least value was recorded. Another research study by He et al.^[61] reported the effect of tillage practices on enzyme activities at various growth stages. Across all the growth stages, enzyme activities of cellobiohydrolase (CBH), β -xylosidase (BXYL), alkaline phosphatase (AP), β -glucosidase (BG), β -N-acetylglucosaminase (NAG) were 17%–169%, 7%–97%, 0.12%–29%, 3%–66%, 23%–137% greater after NT/ST, NT, ST, ST/PT, and PT/NT treatments as compared to plow tillage. The NT/ST treatment resulted in highest soil enzyme activities and yield, and thus was an effective and sustainable method to enhance soil quality and crop production.

Soil microbial biomass carbon

Microbes play a crucial role in controlling different soil functions and soil ecology and microbial community show significant variation across as well as within the landscape. On average, the total biomass of microbes exceeds 500 mg C kg soil⁻¹^[62]. Microbial biomass carbon is an active constituent of SOM which constitutes a fundamental soil quality parameter because SOM serves as a source of energy for microbial processes and is a measure of potential microbial activity^[48,63]. Soil systems that have higher amounts of OM indicate higher levels of MBC. Microbial biomass carbon is influenced by many parameters like OM content in the soil, land use, and

management strategies^[64]. The MBC and soil aggregate stability are strongly related because MBC integrates soil physical and chemical properties responds to anthropogenic activities.

Microbial biomass is regarded as a determinative criterion to assess the functional state of soil. Soils having high functional diversity of microbes which, by and large, occurs under organic agricultural practices, acquire disease and insect-suppressive characteristics that could assist in inducing resistance in plants^[65]. Dou et al.^[66] determined that soil microbial biomass C (SMBC) was 5% to 8% under wheat-based cropping systems and zero tillage significantly enhanced SMBC in the 0–30 cm depth, particularly in the upper 0 to 5 cm. According to Liang et al.^[67], SMBC and soil microbial biomass N (SMBN) in the 0–10 cm surface layer were greater in the fertilized plots in comparison to the unfertilized plots on all sampling dates whereas microbial biomass C and N were highest at the grain filling stage. Mandal et al.^[68] demonstrated that MBC also varied significantly with soil depth. Surface soil possessed a maximum MBC value than lower soil layers due to addition of crop residues and root biomass on the surface soil. The MBC content was highest with combined application of INF along with farmyard manure and GM, whereas untreated plots showed minimum MBC values. The incorporation of CR slows down the rate of mineralization processes; therefore, microbes require more time to decompose the residues and utilize the nutrients released^[69]. On the other hand, incorporation of GR having a narrow C:N ratio enhances microbial activity and consequently accelerates mineralization in the soil. Malviya^[70] also recorded that the SMBC contents were significantly greater under RT than CT, regardless of soil depth which was also assigned to residue incorporation which increases microbial biomass on account of higher carbon substrate in RT.

Naresh et al.^[71] studied the vertical distribution of MBC under no-tillage (NT), shallow (reduced) tillage and normal cultivated fields. A shallow tillage system significantly altered the tillage induced distribution of MBC. In a field experiment, Nakhro & Dkhar^[72] examined the microbial populations and MBC in paddy fields under organic and inorganic farming approaches. The organic source used was a combination of rock phosphate, FYM and neem cake, whereas a mixture of urea, muriate of potash and single super phosphate was used as an inorganic source. The organically treated plots exhibited the highest MBC compared to inorganically treated plots and control. Organic carbon exhibited a direct and significant correlation with bacterial and fungal populations. The addition of organic fertilizers enhanced the content of SOC and consequently resulted in higher microbial count and MBC. Ramdas et al.^[73] investigated the influence of inorganic and organic sources of nutrients (as minerals or INF) applied over a five-year period on SOC, MBC and other variables. It was observed that the addition of FYM and conjoint application of paddy straw (dry) and water hyacinth (PsWh) (fresh) significantly increased the SOC content than vermicompost, *Chromolaena adenophorum* (fresh) and *Glyricidia aculeate* (fresh), and *Sesbania rostrata* (fresh).

Xu et al.^[74] evaluated the influence of long-term fertilization strategies on the SOC content, soil MBN, soil MBC, and soil microbial quotient (SMQ) in a continuous rice system and observed that MBC at the main growth stages of early and late rice under 30% organic matter and 70% mineral fertilizer and 60% organic matter and 40% mineral fertilizer treatments was greater as compared to mineral fertilizer alone (MF), rice straw

residues and mineral fertilizer (RF), and no fertilizer (CK) treatments. However, SMBC levels at late growth stages were greater in comparison to early growth stages. A recent study by Xiao et al.^[75] demonstrated that increasing tillage frequency (no-tillage, semi-annual tillage, and tillage after every four months, two months, and one month) decreased soil MBC. Microbial biomass carbon content was significantly greater in no-till treatment (597 g kg⁻¹) than in tillage every four months (421 g kg⁻¹), two months (342 g kg⁻¹) and one month (222 g kg⁻¹). The decrease in the content of MBC in association with tillage practices is due to soil perturbations which enhanced soil temperature, diminished soil moisture content, and resulted in the destruction of microbial habitat and fungal hyphae. Therefore, the MBC content eventually affected the N cycle.

Li et al.^[76] reported that in comparison to CT, NT and RT resulted in increased MBC content and NT significantly increased MBC by 33.1% over CT. Furthermore, MBC concentration was 34.1% greater in NT than RT. The increase in MBC concentration was correlated with the results of increase in SOC concentration. Site-specific factors including soil depth and mean annual temperature significantly affected the response ratio of MBC under NT as compared to the duration of NT.

Soil microbial biomass nitrogen

Microbial biomass nitrogen (MBN) is a prominent indicator of soil fertility as it quantifies the biological status of soil. Soil MBN is strongly associated with organic matter of the soil. The nitrogen in MBN has a rapid turnover rate thereby reflecting the changes in management strategies way before the transformations in total N are discernable^[77].

In an experiment on continuous silage maize cultivation with crop rotation, Cerny et al.^[78] observed that organic fertilizers exerted an affirmative influence on the soil MBN. During the application of organic manure MBN decreased, but there was higher MBN content as compared to control. However, addition of mineral nitrogenous fertilizers exerted an adverse effect on MBN content in experiments with maize. El-Sharkawi^[79] recorded that organic matter-treated pots resulted in maximum MBN content than urea-treated pots. The sludge application enhanced total MBN and, therefore, could

implicitly benefit crop production particularly in poor soils^[18]. Sugihara et al.^[80] observed that during the grain-filling stage in maize, residue and/or fertilizer addition exerted a pronounced influence on soil microbial dynamics; however, a clear effect of residue and /or fertilizer addition was not observed. Microbial biomass nitrogen reduced dramatically from 63–71 to 18–33 kg N ha⁻¹ and C:N ratio at the same time increased more than ten-fold in all plots.

Malik et al.^[81] apprised that the organic amendments significantly enhanced MBN concentrations up to 50% more than the unamended soil. Wang et al.^[82] evaluated the influence of organic materials on MBN content in an incubation and pot experiment with acidic and calcareous soils. The results revealed that MBN content which was affected by the different forms of organic amendments, increased by 23.37%–150.08% and 35.02%–160.02% in acidic and calcareous soils, respectively. The MBN content of both soils decreased with the increase in the C/N ratio of the organic materials, though a higher C/N ratio was effective for sustaining a greater MBN content for a very long time.

Dhaliwal & Bijay-Singh^[52] observed higher MBN levels in NT soils (116 kg ha⁻¹) than in cultivated soils (80 kg ha⁻¹). Kumar et al.^[83] ascertained that in surface layer, MBN content was 11.8 mg kg⁻¹ in CT which increased to 14.1 and 14.4 mg kg⁻¹ in ZT and RT without residue retention and 20.2, 19.1 and 18.2 mg kg⁻¹ in ZT, RT and CT with residue incorporation, respectively (Table 1). In the subsurface layer, the increased tendency on account of tillage and crop residue retention was identical to those of 0–15 cm layer but the magnitude was comparatively meagre (Table 1). In comparison to control, the persistent retention of crop residues led to significant accrual of MBN in the surface layer.

Xiao et al.^[75] determined that the MBN content decreased with tillage treatment having highest value in no tillage treatment, however, the difference among the treatments was negligible. Soil perturbations decreased the aggregate size and thus lower the soil aeration and exposure of fresh organic matter which restricted the growth of microorganisms. The results also concluded that MBN content is highly sensitive to tillage. Ginakes et al.^[84] assessed the impact of zone tillage

Table 1. Effect of different treatments on contents of various fractions of soil organic carbon^[38].

Treatments	PMN (mg kg ⁻¹)		MBC (mg kg ⁻¹)		MBN (mg kg ⁻¹)		DOC (mg kg ⁻¹)	
	Depths (cm)							
	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30
Tillage practices								
ZTR	12.4	11.2	562.5	471.1	20.2	18.9	198.6	183.6
ZTWR	8.5	7.6	350.4	302.1	14.1	12.6	167.1	159.2
RTR	10.6	9.9	490.2	399.3	19.1	17.2	186.4	171.6
RTWR	7.6	6.6	318.1	299.8	14.4	13.7	159.5	148.7
CTR	9.3	8.5	402.9	354.4	18.2	16.6	175.9	168.9
CT	6.7	5.6	307.9	289.5	11.8	9.7	142.5	134.6
Nitrogen management								
Control	3.6	2.8	218.3	202.9	10.8	10.4	103.7	92.3
80 kg N ha ⁻¹	5.3	4.4	241.1	199.4	14.9	12.2	128.3	116.9
120 kg N ha ⁻¹	8.9	7.6	282.7	220.9	16.5	16.1	136.8	123.6
160 kg N ha ⁻¹	9.8	8.4	343.9	262.9	19.4	18.1	164.8	148.9
200 kg N ha ⁻¹	10.4	9.7	346.3	269.6	22.7	21.7	155.7	136.4

ZTR = Zero tillage with residue retention, ZTWR = Zero tillage without residue retention; RTR = Reduced tillage with residue retention, RTWR = Reduced tillage without residue retention, CTR = Conventional tillage with residue incorporation; CT = Conventional tillage without residue incorporation.

intensity on MBN in a corn-kura clover cropping sequence. Microbial biomass nitrogen was influenced by time and type of tillage treatment. Temporal studies revealed that MBN was higher after tillage treatment than the values possessed before tillage. Under different tillage treatments, higher values were recorded in ST (shank-till) and DT (double-till) over NT and RZT (zone-till) treatments.

Potentially mineralizable nitrogen in the soil

Another biological parameter, PMN, is a crucial parameter of soil fertility due to its association with soil N supply for crop growth. Also, PMN indicates the status of soil microbial community associated with PMN, whether it is improving or degrading. Forest soils are characterized by greater levels of PMN than CT receiving conventional chemical fertilizers which could be assignable to improved microbial activity in the former soils than the latter^[48,77]. Aulakh et al.^[85] assessed the effect of various combinations of fertilizer N, P, FYM and wheat residue (WR) applied to soybean and soybean residues added to wheat under CT and CA. The added fertilizers of N and P, FYM, and crop residue enhanced the mean weight diameter and water-stable aggregates thus favoured the development of macro-aggregates. The treatment INF + FYM + crop residue performed better among all the treatments. The net flux of mineral nitrogen from the mineralizable fraction is used to measure potentially mineralizable N which indicates the balance between mineralization and immobilization by soil microbes^[77]. Nitrogen mineralization is widely used to assess the ability of SOM to supply inorganic nitrogen in the form of nitrate which is the most common form of plant-available nitrogen. Kumar et al.^[83] observed an increase in PMN which was higher in surface soil than sub-surface soil thereby implying that high OC accumulation on account of crop residue retention was the most probable cause.

Verma & Goyal^[86] assessed the effect of INM and organic manuring on PMN and observed that PMN was substantially affected by different organic amendments. Potentially mineralizable nitrogen varied between 19.6–41.5 mg kg⁻¹ soil with greater quantity (2.5%) in vermicompost applied plots than FYM treated plots. The INF treatments resulted in lower PMN content which could be due to nutrient immobilization by microbes. Mahal et al.^[87] reported that no-till resulted in higher PMN content than conventional tillage treatments. This trend was due to the maintenance of SOM due to the residue cover and reduction of soil erosion under no-tillage system^[88]. On the contrary, tillage practices led to the loss of SOC owing to loosened surface soil and higher mineralization of SOM.

Microbial respiration in soil

Soil respiration is referred as the sum of CO₂ evolution from intact soils because of the respiration by soil organisms, mycorrhizae and roots^[89]. Various researchers have proposed soil respiration as a potential indicator of soil microbial activity^[52,77]. Gilani & Bahmanyar^[90] observed that addition of organic amendments enhanced soil respiration more than the control and synthetic fertilizer treatments. Moreover, among organic amendment treatments, highest soil respiration was observed in sewage-sludge treated soils. Under controlled conditions in saline-sodic soil, Celis et al.^[91] reported that sewage sludge resulted in a higher soil respiration rate than mined gypsum and synthetic gypsum. The application of gypsum because of minimal organic matter intake had little effect on soil respiration. The addition of organic matter especially during early

spring led to higher microbial biomass and soil respiration albeit diminished levels of nitrate-N. Moreover, SOM hinders the leaching of nitrate ions thereby resulting in a better soil chemical environment^[71].

Faust et al.^[92] observed that microbial respiration was associated with volumetric water content. The respiration declined with less availability of water, thus the lesser the tillage intensity, the more the volumetric water content which consequently resulted in higher microbial respiration. Another study by Bongiorno et al.^[93] reflected the influence of soil management intensity on soil respiration. Reduced tillage practices resulted in 51% higher basal respiration than CT. Furthermore, this investigation suggested that microbial catabolic profile could be used as a useful biological soil quality indicator. Recently, Kalkhajeh et al.^[94] ascertained the impact of simultaneous addition of N fertilizer and straw-decomposing microbial inoculant (SDMI) on soil respiration. The SDMI application boosted the soil microbial respiration which accelerated the decomposition of straw due to N fertilization. The C/N ratio did not affect the microbial respiration at elongation and heading stages, whereas N fertilization enhanced the microbial respiration to a greater extent than the unfertilized control. Additionally, the interaction between sampling time and basal N application significantly affected microbial respiration.

Gong et al.^[95] appraised the effect of conventional rotary tillage and deep ploughing on soil respiration in winter wheat and observed that deep ploughing resulted in a higher soil respiration rate than conventional rotary tillage. Soil moisture content and temperature are the dominating agents influencing soil respiration which is restricted by the soil porosity.

Effect of fertilization and tillage practices on soil organic carbon (SOC) transformations

Soil organic carbon plays a vital role in regulating various soil functions and ecosystem services. It is influenced by numerous factors like tillage practices and fertilization. Moreover, modified management practices may prove beneficial to avoid SOC loss by increasing its content. An exogenous supply of fertilizers may alter the chemical conditions of soil and thus result in transformation of SOC. Tillage practices lead to frequent soil disturbances which reduce the size of soil aggregates and accelerate the oxidation of SOC thereby reducing its content. The literature on the influence of fertilization and tillage practices on the transformation of SOC is discussed below.

Build-up of soil organic carbon (SOC) in soil

Soil organic carbon is a major part of the global carbon cycle which is associated not only with the soil but also takes part in the C cycling through vegetation, oceans and the atmosphere (Figs 1 & 2). Soil acts as a sink of approximately 1,500 Pg of C up to 1 m depth, which is greater than its storage in the atmosphere (approximately 800 Pg C) and terrestrial vegetation (500 Pg C) combined^[96]. This dynamic carbon reservoir is continuously cycling in diverse molecular forms between the different carbon pools^[97]. Fertilization (both organic and mineral) is one of the crucial factors that impart a notable influence on OC accretion in the soil. Many researchers have studied the soil C dynamics under different fertilizer treatments. Though inorganic fertilizers possess the advantage of easy handling, application and storage, they do not contribute to soil organic carbon. On the contrary, regardless of management method, plant residues are known to increase organic carbon content.

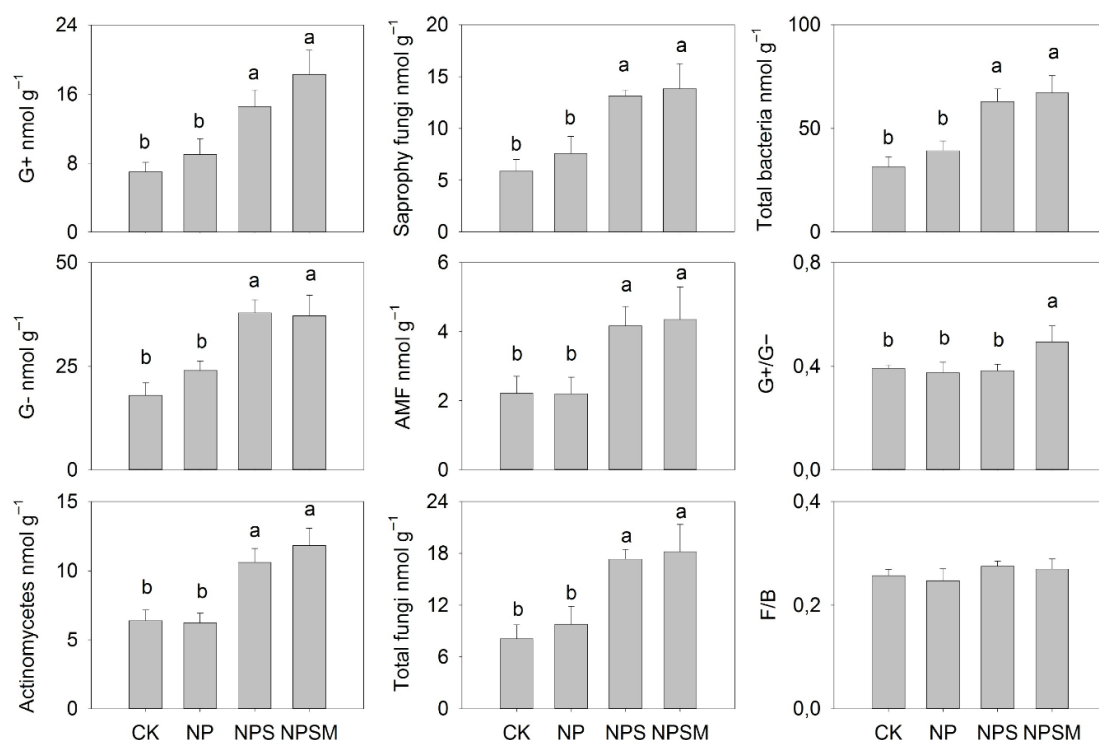


Fig. 1 Impact of different fertilization regimes on abundance of the microbial biomarker groups. Error bars represent the standard error of the means and different letters indicate significant differences at $p < 0.05$ among treatments. Source: Li et al.^[60].

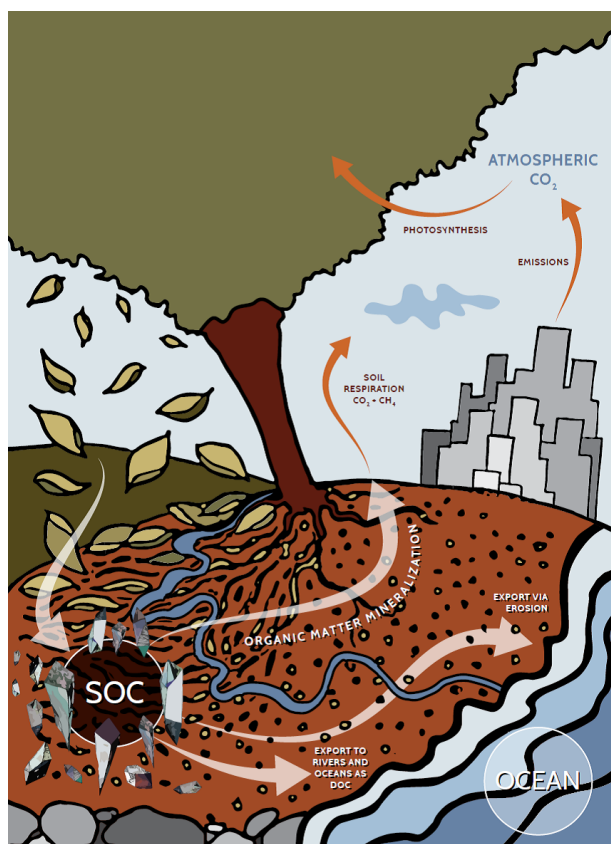


Fig. 2 Soil organic carbon (SOC) dynamics in the global carbon cycle.

Katkar et al.^[98] reported a higher soil quality index under conjunctive nutrient management strategies comprising addition of compost and green leaves along with mineral nutrients. Mazumdar et al.^[99] investigated the impact of crop residue (CR), FYM, and leguminous green manure (GM) on SOC in continuous rice-wheat cropping sequence over a 25-year period. At the surface layer, the maximum SOC content was recorded under NPK + FYM than NPK + CR and NPK + GM treatments. SOC was significantly lower under sole application of INFs (NPK) than the mixed application of organic and inorganic treatments. A higher range of SOC content was recorded at a depth of 0.6 m in the rice-wheat system (1.8–6.2 g kg⁻¹) in farmyard manure (FYM)-treated plots than 1.7–5.3 g kg⁻¹ under NPK, and 0.9–3.0 g kg⁻¹ in case of unfertilized plots^[100]. In a research study Dutta et al.^[101] reported that rice residue had a higher decomposition rate ($k = 0.121$ and 0.076 day⁻¹) followed by wheat (0.073 and 0.042 day⁻¹) and maize residues (0.041 day⁻¹) when their respective residues placed on soil surface than incorporated in the soils. Naresh et al.^[102] found FYM and dhaincha as GM/sulphitation press mud (SPM) treatments are potent enough to enhance the SOC. Maximum SOC content was noted in 0–5 cm depth that reduced gradually along the profile. In surface soil, the total organic content (TOC) under different treatments varied with source used to supply a recommended dose of nitrogen (RDN) along with conventional fertilizer (CF).

Cai et al.^[103] ascertained that long-term manure application significantly improved SOC content in different size fractions which followed the sequence: 2,000–250 μ m > 250–53 μ m > 53 μ m fraction. Naresh et al.^[22] determined that mean SOC content increased from 0.54% in control to 0.65% in RDF and 0.82% in RDF + FYM treatment and improved enzyme activity; thus, ultimately influenced nutrient dynamics under field conditions. The treatments RDF + FYM and NPK resulted in 0.28

Mg C ha⁻¹ yr⁻¹ and 0.13 Mg C ha⁻¹ yr⁻¹, respectively and thus higher sequestration than control. Zhao et al.^[104] determined that in the surface layer, significant increase in SOC content in each soil aggregate was noticed under straw incorporation treatments over no straw incorporated treatments (Fig. 3). Moreover, the aggregate associated OC was significantly higher in the surface layer than the sub-surface layer. The highest increment in aggregate-associated OC was noted in both maize and wheat straw (MR-WR) added plots followed by MR and least in WR. Besides, all of the three straw-incorporated treatments exhibited notable increase in SOC stock in each aggregate fraction in the surface layer of the soil. In the subsurface (20–40 cm) layer under MR-WR, significant rise in SOC stock of small macro-aggregates was observed, whereas there was a reduction in SOC stock in the silt + clay fraction than other treatments. The straw-incorporated treatments increased the quantity of mineral-associated organic matter (mSOM) and intra-aggregate particulate organic matter (iPOM) within small macro-aggregates and micro-aggregates especially in the topmost layer of the soil.

Srinivasarao et al.^[105] reported that SOC content was reduced with the addition of INFs (100% RDN) alone as compared to the conjunctive application of inorganic and organic or sole FYM

treatments. Earlier, Srinivasarao et al.^[106] reported that FYM treated plots exhibited greater per cent increase in SOC stock than mineral fertilized plots and control. Tong et al.^[107] ascertained that the application of NP and NPK significantly improved SOC stocks. On the contrary, fertilized soils could also exhibit decrease in carbon content than control. Naresh et al.^[108] determined that higher biomass C input significantly resulted in greater particulate organic carbon (POC) content. Zhang et al.^[109] ascertained that long-term addition of NPK and animal manures significantly improved SOC stocks by a magnitude of 32%–87% whereas NPK and wheat/ and or maize straw incorporation enhanced the C stocks by 26%–38% than control. Kamp et al.^[110] determined that continuous cultivation without fertilization decreased SOC content by 14% than uncultivated soil. However, super optimum dose of NPK, balanced NPK fertilization and integration of NPK with FYM not only improved SOC content but also SOC stocks over the first year. In conventionally tilled cotton-growing soils of southern USA, Franzluebbers et al.^[111] estimated that carbon sequestration averaged 0.31 ± 0.19 Mg C ha⁻¹ yr⁻¹. Mandal et al.^[112] reported maximum SOC stock in the surface layer of the soil (0–15 cm) which progressively diminished with depth in each land use system. A significant decrease in SOC stock along the profile

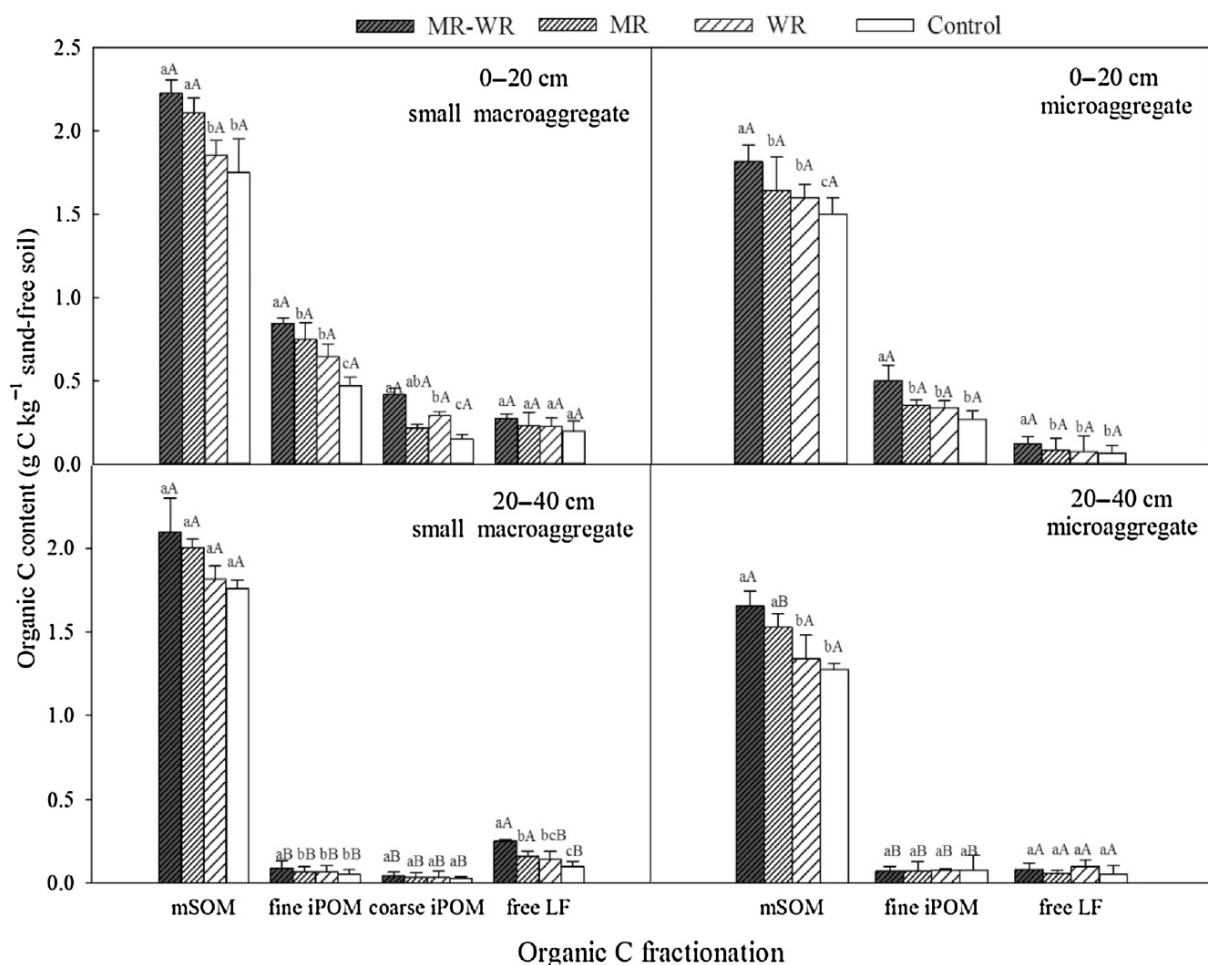


Fig. 3 Distribution of OC in coarse iPOM (intra-aggregate particulate organic matter) fine iPOM, mSOM (mineral-associated matter), and free LF (free light fraction) of small macro-aggregates and micro-aggregates in the 0–20 cm and 20–40 cm soil layers under MR-WR (return of both maize and wheat straw), MR (maize straw return), WR (wheat straw return). Different lowercase and uppercase letters indicate significant differences at $p < 0.05$ among treatments and depths respectively^[104].

depth was also observed by Dhaliwal et al.^[47] in both crop-lands and agroforestry. In the topmost soil layer, highest SOC stock was recorded in rice–fallow system while the lowest was in the guava orchard^[112].

Nath et al.^[113] determined that there was accrual of higher TOC in surface layers as compared to lower layers of soil under paddy cultivation. This accrual could be added to left-over crop residues and remnant root biomass which exhibited a decreasing trend with soil depth. Das et al.^[114] determined that integrated use of fertilizers and organic sources resulted in greater TOC as compared to control or sole fertilizer application. Fang et al.^[115] observed that the cumulative carbon mineralization differed with aggregate size in top soils of broad-leaved forests (BF) and coniferous forests (CF). However, in deep soil it was greater in macro-aggregates as compared to micro-aggregates in BF but not in CF (Fig. 4). By and large, the percent SOC mineralized was greater in macro-aggregates as compared to micro-aggregates. Dhaliwal et al.^[100] ascertained that SOC accrual was considerably influenced by residue levels and tillage in surface soil (0–20 cm); albeit no variation was observed at lower depth (20–40 cm). The SOC content was greater in zero-tilled and permanently raised beds incorporated with residues as compared to puddled transplanted rice

and conventionally planted wheat. Pandey et al.^[116] reported that no-tillage prior to sowing of rice and wheat increased soil organic carbon by $0.6 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$. The carbon sequestration rate on account of no-tillage or reduced tillage ranged between $0\text{--}2,114 \text{ kg ha}^{-1} \text{ yr}^{-1}$ in the predominant cropping system of South Asia, Xue et al.^[117] observed that the long-term conventional tillage, by and large, exhibited a significant decline in SOC owing to degradation of soil structure, exposing protected soil organic matter (intra-soil aggregates) to microbes. Therefore, the adoption of no-tillage could hamper the loss of SOC thereby resulting in a greater or equivalent quantity of carbon in comparison to CT (Fig. 5).

Singh et al.^[118] determined that carbon stock in the 0–40 cm layer increased by 39, 35 and 19% in zero-tilled clay loam, loam, and sandy loam soils, respectively as compared to conventional tilled soils over a period of 15 years. Kuhn et al.^[119] also apprised about the advantages of NT over CT vis-a-vis SOC stocks across soil depths. In the surface layer (0–20 cm) NT, by and large, resulted in higher SOC stocks as compared to CT; however, SOC stocks exhibited a declining trend with soil depth, in fact, became negative at depths lower than 20 cm. Sapkota et al.^[120] observed that over a period of seven years, direct dry-seeded rice proceeded by wheat cultivation with

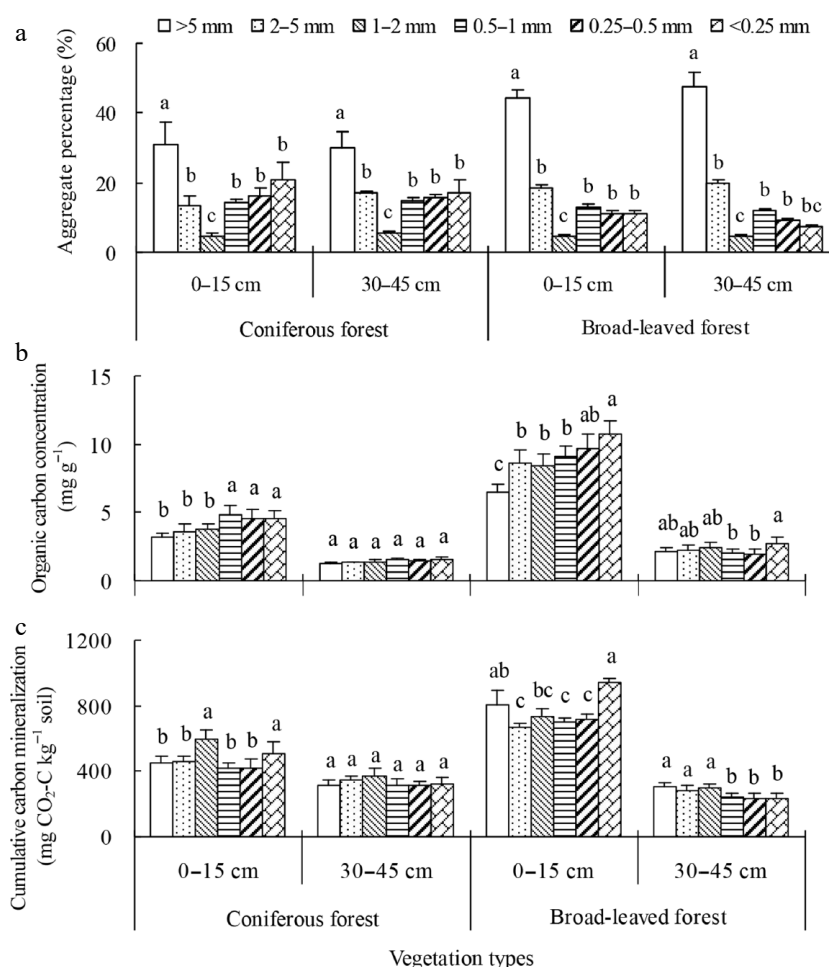


Fig. 4 (a) Soil aggregate fractions of two depths in two restored plantations of subtropical China, (b) organic carbon and (c) its mineralization from various soil aggregates within 71 d at various soil depths in two restored plantations of subtropical China. Error bars show the standard error of the mean. The different letters represent significant differences among the different soil aggregate fractions within a depth at $p < 0.05$ ^[115].

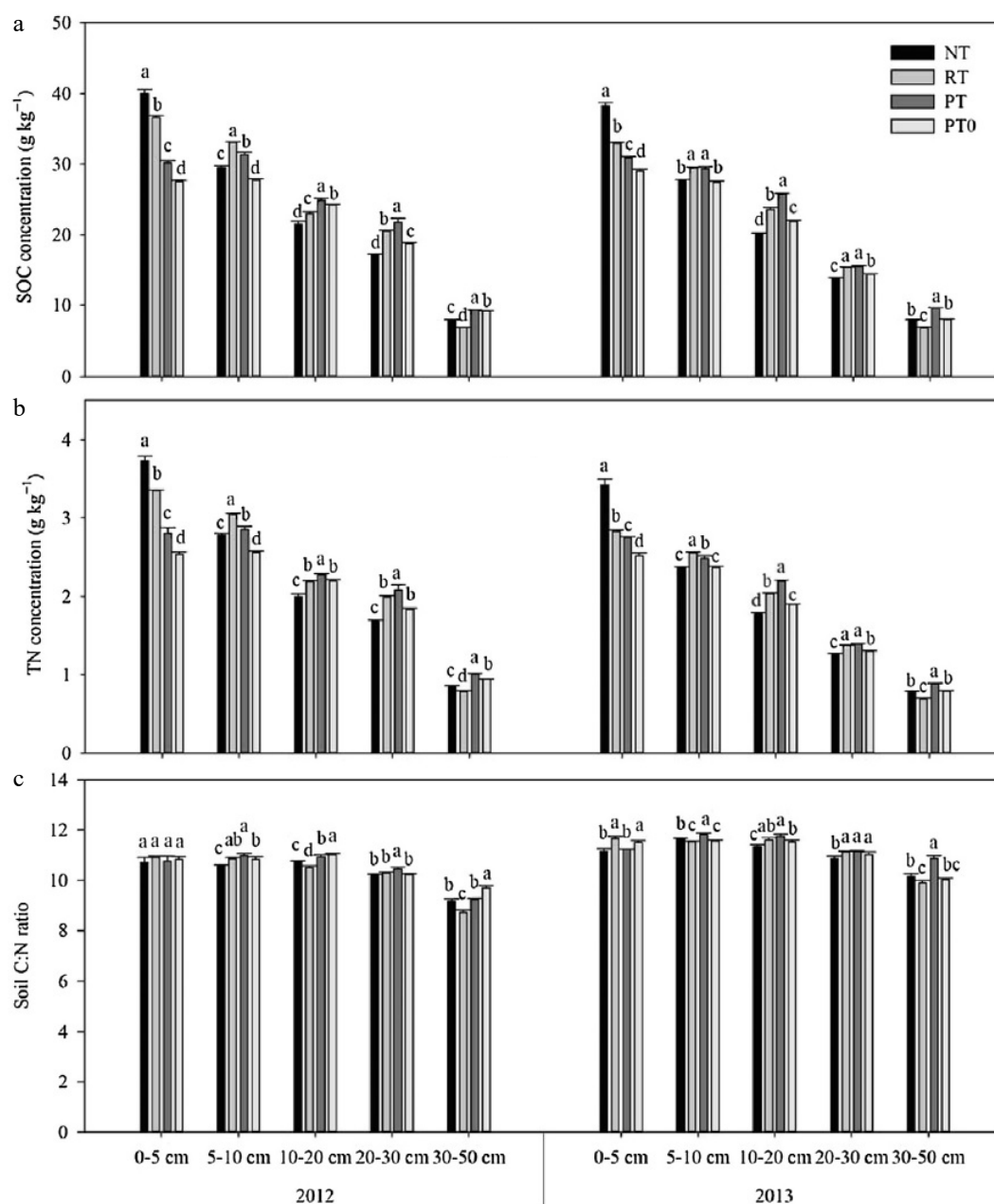


Fig. 5 The concentrations of (a) SOC, (b) total nitrogen (TN), and (c) soil C:N ratio for 0–50 cm profile under different tillage treatments in 2012 and 2013. NT = no-till with residue retention; RT = rotary tillage with residue incorporation; PT = plow tillage with residue incorporation; and PT0 = plow tillage with residue removed. The lowercase letters indicate statistical difference among treatments at $p < 0.05$ [117].

residue retention enhanced SOC at 0–60 cm depth by a magnitude of 4.7 and 3.0 t C ha⁻¹ in zero-tillage (ZTDSR-ZTW + R) and without tillage (PBDSR-PBW + R), respectively. On the contrary, the conventional tillage rice-wheat cropping system (CTR-CTW) decreased the SOC up to 0.9 t C ha⁻¹ (Table 2).

Labile carbon (LC) fractions in soil

Labile organic carbon (LC) is that fraction of SOC that is rapidly degraded by soil microbes, therefore, having the highest turnover rate. This fraction can turn over quickly on account of the change in land use and management strategies. From the crop production perspective, this fraction is crucial as it sustains the soil food cycle and, hence, considerably impacts

nutrient cycling thereby altering soil quality and productivity. Short-term management could influence the labile fraction of carbon [121]. However, some site-specific problems and regional factors may influence their distribution in soil layers [102].

Banger et al. [122] observed significant alteration in labile pools of C, for instance, particulate organic matter (POM), water-soluble C (WSC) and light fraction of C (LFC) because of the addition of fertilizers and/or FYM over a 16-year period. Particulate organic matter, LFC and WSC contributed 24%–35%, 12%–14% and 0.6%–0.8%, respectively, towards SOC. The increase in concentration of SOC including its pools like POC and the sequestration rate due to integrated nutrient management was also reported by Nayak et al. [123]. Gu et al. [124] observed that

Table 2. Influence of tillage and crop establishment methods on SOC stock and its temporal variation under rice–wheat system^[120].

Tillage and crop establishment methods	Depths (m)				
	0–0.05	0.05–0.15	0.15–0.3	0.3–0.6	0–0.6
Total SOC (t/ha)					
CTR-CTW	3.5 ^e	7.1 ^c	8.7	7.0	26.2 ^c
CTR-ZTW	3.9 ^d	7.6 ^{bc}	8.8	6.5	26.7 ^c
ZTDSR-CTW	4.2 ^d	7.5 ^{bc}	9.2	6.3	27.3 ^c
ZTDSR-ZTW	4.9 ^c	8.9 ^{ab}	8.2	6.2	28.2 ^{bc}
ZTDSR-ZTW+R	6.1 ^a	9.0 ^{ab}	9.8	6.8	31.8 ^a
PBDSR-PBW+R	5.5 ^b	9.3 ^a	9.3	6.0	30.1 ^{ab}
MSD	0.4	1.7	2.0	1.4	2.49
Treatment effect (p value)	<0.001	0.04	0.158	0.267	<0.001
Initial SOC content	3.6 ± 0.15	8.1 ± 1.39	8.78 ± 1.07	6.7 ± 0.73	27.1 ± 1.21
Change in SOC over seven years (t/ha)					
CTR-CTW	−0.16	−0.99	−0.04	0.28	−0.90
CTR-ZTW	0.28	−0.50	0.01	−0.20	−0.41
ZTDSR-CTW	0.62	−0.57	0.45	−0.34	0.16
ZTDSR-ZTW	1.34	0.84	−0.62	−0.46	1.09
ZTDSR-ZTW+R	2.49	0.96	1.04	0.16	4.66
PBDSR-PBW+R	1.89	1.22	0.51	−0.64	2.98

CTR-CTW = Conventionally tilled puddled transplanted rice followed by conventionally tilled wheat, CTR-ZTW = Conventionally tilled puddled transplanted rice followed by zero-tilled wheat, ZTDSR-CTW = Zero-tilled direct dry-seeded rice followed by conventionally tilled wheat, ZTDSR-ZTW = Zero-tilled direct dry-seeded rice followed by zero-tilled wheat, ZTDSR-ZTW+R = Zero-tilled direct dry-seeded rice followed by zero-tilled wheat with residue retention, PBDSR-PBW+R = Direct dry-seeded rice followed by direct drilling of wheat both on permanent beds with residue retention, MSD, minimum significant difference. Significant different letters indicate significant differences at $p < 0.05$.

mulch-treated soils (straw and grass mulch) had significantly greater levels of LOC, POC, DOC and EOC as compared to no mulch-treated soils which could be adduced to the addition of straw, root and its sections into the soil. The content of labile C fractions across all treatments exhibited a decreasing trend with soil depth^[23, 102, 125].

In a long-term experiment, Anantha et al.^[126] observed that the total organic carbon apportioned into labile carbon, non-labile, less labile, and very labile carbon constituted around 18.7%, 19.3%, 20.6% and 41.4% of the TOC, respectively (Table 3). Zhu et al.^[20] determined that straw incorporation had a substantial impact on TOC and labile C fractions of the soil which were greater in straw incorporated treatments as compared to non-straw treatments across all the depths. Wang et al.^[127] reported that the light fraction organic carbon (LFOC) and DOC were significantly greater in the straw-applied treatments than the control by a magnitude of 7%–129% for both the early and late season rice. The treatments NPK + FYM or NPK + GR + FYM resulted in greater content of very labile and labile C fractions whereas non-labile and less labile fractions were greater in control and NPK + CR treatment. There was 40.5% and 16.2% higher C build-up in sole FYM treated plots and 100% NPK + FYM, respectively over control. On the other hand, a net depletion of 1.2 and 1.8 Mg ha^{−1} in carbon stock was recorded under 50% NPK and control treatments, respectively. Out of the total C added through FYM, only 28.9% was stabilized as SOC, though an external supply of OM is a significant source of soil organic carbon^[69]. Hence, to sustain the optimum SOC level at least an input of 2.3 Mg C ha^{−1} y^{−1} is required. A comparatively greater quantity of soil C in passive

Table 3. Oxidisable organic carbon fractions in soils (g kg^{−1}) at different layers^[126].

Treatment	Depths (cm)			
	0–15	15–30	30–45	Total
Very Labile C				
Control	3.6 ± 0.5 ^c	1.4 ± 0.3 ^b	1.3 ± 0.2 ^a	6.3 ± 0.4 ^b
50% NPK	4.6 ± 0.3 ^{bc}	2.1 ± 0.7 ^{ab}	1.5 ± 0.1 ^a	8.1 ± 0.9 ^a
100% NPK	4.4 ± 0.3 ^{bc}	2.3 ± 0.2 ^a	1.4 ± 0.5 ^a	8.0 ± 0.7 ^a
150% NPK	5.0 ± 0.2 ^{ab}	2.6 ± 0.2 ^a	1.5 ± 0.1 ^a	9.0 ± 0.3 ^a
100% NPK + FYM	4.8 ± 0.2 ^{ab}	2.0 ± 0.2 ^{ab}	1.3 ± 0.3 ^a	8.1 ± 0.2 ^a
FYM	5.9 ± 1.3 ^a	2.2 ± 0.2 ^a	1.4 ± 0.3 ^a	9.5 ± 1.6 ^a
Fallow	4.2 ± 0.7 ^{bc}	1.5 ± 0.5 ^b	0.7 ± 0.3 ^b	6.3 ± 0.8 ^b
Labile C				
Control	2.4 ± 0.3 ^a	1.0 ± 0.2 ^a	0.8 ± 0.4 ^a	4.2 ± 0.6 ^a
50% NPK	1.7 ± 0.4 ^{ab}	0.9 ± 0.5 ^a	0.7 ± 0.2 ^a	3.3 ± 0.7 ^a
100% NPK	1.8 ± 0.4 ^{ab}	0.8 ± 0.5 ^a	0.6 ± 0.3 ^a	3.2 ± 0.8 ^a
150% NPK	1.2 ± 0.3 ^b	0.7 ± 0.2 ^a	0.9 ± 0.2 ^a	2.8 ± 0.4 ^a
100% NPK + FYM	1.9 ± 0.3 ^{ab}	0.7 ± 0.2 ^a	0.7 ± 0.3 ^a	3.4 ± 0.2 ^a
FYM	2.5 ± 0.9 ^a	0.7 ± 0.3 ^a	0.7 ± 0.2 ^a	3.9 ± 0.9 ^a
Fallow	2.2 ± 1.0 ^{ab}	1.0 ± 0.3 ^a	1.0 ± 0.4 ^a	4.1 ± 1.1 ^a
Less labile C				
Control	1.5 ± 0.3 ^c	0.6 ± 0.4 ^c	0.4 ± 0.0 ^c	2.6 ± 0.7 ^d
50% NPK	1.8 ± 0.1 ^c	0.4 ± 0.1 ^c	0.5 ± 0.2 ^c	2.7 ± 0.1 ^{cd}
100% NPK	2.5 ± 0.3 ^{ab}	0.8 ± 0.1 ^{bc}	1.1 ± 0.2 ^{ab}	4.4 ± 0.1 ^b
150% NPK	2.6 ± 0.2 ^a	0.9 ± 0.1 ^{bc}	0.4 ± 0.2 ^c	3.9 ± 0.1 ^b
100% NPK + FYM	2.7 ± 0.6 ^a	1.5 ± 0.2 ^a	1.4 ± 0.1 ^a	5.6 ± 0.7 ^a
FYM	1.9 ± 0.7 ^{bc}	1.7 ± 0.2 ^a	1.0 ± 0.2 ^b	4.5 ± 0.7 ^{ab}
Fallow	1.5 ± 0.3 ^c	1.3 ± 0.7 ^{ab}	0.9 ± 0.4 ^b	3.8 ± 1.2 ^{bc}
Non labile C				
Control	1.2 ± 0.5 ^b	1.2 ± 0.3 ^a	0.2 ± 0.2 ^b	2.6 ± 0.5 ^b
50% NPK	1.2 ± 0.9 ^b	1.7 ± 0.8 ^a	0.7 ± 0.4 ^{ab}	3.5 ± 1.8 ^{ab}
100% NPK	1.3 ± 0.6 ^b	1.5 ± 0.6 ^a	0.5 ± 0.2 ^{ab}	3.3 ± 1.0 ^{ab}
150% NPK	1.4 ± 0.3 ^b	1.5 ± 0.2 ^a	0.8 ± 0.1 ^a	3.7 ± 0.3 ^{ab}
100% NPK + FYM	2.0 ± 0.8 ^b	1.3 ± 0.1 ^a	0.3 ± 0.3 ^{ab}	3.5 ± 0.7 ^{ab}
FYM	3.7 ± 1.3 ^a	1.0 ± 0.2 ^a	0.5 ± 0.5 ^{ab}	5.1 ± 1.9 ^a
Fallow	2.1 ± 0.2 ^b	1.4 ± 0.7 ^a	0.4 ± 0.2 ^{ab}	3.9 ± 0.9 ^{ab}

Values in the same column followed by different letters are significantly different at $p < 0.001$, ± indicates the standard deviation values of the means.

pools was observed in 100% NPK + FYM treatment. The increase in allocation of C into the passive pool was about 33%, 35%, 41% and 39% of TOC in control, suboptimal dose, optimal dose and super optimal dose of NPK which indicates that the concentration of passive pools increased with an increase in fertilization doses. Water-soluble carbon (WSC) was 5.48% greater in the upper soil layer as compared to lower layer of soil. In surface soil (0–15 cm), the values of light fraction carbon (LFC) were 81.3, 107.8, 155.2, 95.7, 128.8, 177.8 and 52.7 mg kg^{−1} in ZT without residue retention, ZT with 4 t ha^{−1} residue retention, ZT with 6 t ha^{−1} residue retention, FIRB without residue addition and FIRB with 4 and 6 t ha^{−1} residue addition and CT, respectively (Table 4). Tiwari et al.^[128] determined that the decrease in POC was due to reduction in fine particulate organic matter in topsoil whereas decrement in dissolved organic carbon was observed largely in subsoil. Therefore, in surface soils fine POC and LFOC might be regarded as preliminary evidence of organic C alteration more precisely, while DOC could be considered as a useful indicator for subsoil. Reduction in allocations of fine POC, LFOC and DOC to SOC caused by tillage and straw management strategies indicated the decline in quality of SOC. A higher SOC concentration was recorded in

Table 4. Influence of tillage and nitrogen management on distribution of carbon fractions in soil^[83].

Treatments	WSOC (g kg ⁻¹)		SOC (g kg ⁻¹)		OC (g kg ⁻¹)		BC (g kg ⁻¹)		POC (mg kg ⁻¹)		PON (mg kg ⁻¹)		LFOC (mg kg ⁻¹)		LFON (mg kg ⁻¹)	
	Depths (cm)															
	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30	0–15	15–30
Tillage practices																
ZTR	28.8	26.2	23.1	19.3	9.61	9.13	4.69	4.28	1342.8	967.9	119.5	108.1	194.7	154.8	14.8	12.3
ZTWR	25.3	24.6	18.4	14.8	7.87	7.21	3.76	3.19	981.1	667.4	94.6	86.5	120.5	104.7	11.8	10.3
RTR	27.0	25.9	22.4	18.2	8.68	8.17	4.13	3.87	1230.2	836.9	109.7	97.8	170.9	144.9	13.7	11.6
RTWR	23.7	21.8	18.1	14.2	7.66	7.07	3.12	2.96	869.4	604.4	82.6	76.6	107.1	97.3	9.7	8.6
CTR	26.1	24.4	21.8	17.4	8.49	7.96	3.82	3.48	1099.1	779.4	98.4	89.3	143.8	115.9	12.8	10.9
CT	21.8	20.9	16.1	13.1	6.21	5.64	2.89	2.63	617.5	481.8	69.2	57.6	90.8	73.6	9.6	7.9
Nitrogen management																
Control	21.1	14.9	16.1	13.1	6.13	5.48	1.58	1.07	709.7	658.6	31.7	26.3	123.9	104.3	6.4	5.8
80 kg N ha ⁻¹	28.3	21.2	17.8	14.7	6.46	6.16	2.46	1.75	860.7	785.6	68.4	56.2	132.8	116.1	7.6	6.9
120 kg N ha ⁻¹	29.5	22.1	19.1	16.1	7.25	6.71	3.26	2.18	952.2	808.9	89.5	78.5	150.6	127.6	9.7	8.6
160 kg N ha ⁻¹	30.2	23.1	20.8	18.2	7.75	7.28	3.82	2.66	1099.5	823.8	96.8	83.4	168.5	145.7	10.2	9.8
200 kg N ha ⁻¹	31.1	25.4	21.3	18.7	7.93	7.48	4.15	3.42	1153.1	898.4	103.9	97.3	176.2	152.9	11.7	10.6

WSOC = Water soluble organic carbon, SOC = Total soil organic carbon, OC = Oxidizable organic carbon, BC = Black carbon, POC = particulate organic carbon, PON = particulate organic nitrogen, LFOC = labile fraction organic carbon, and LFON = labile fraction organic nitrogen. ZTR = Zero tillage with residue retention, ZTWR = Zero tillage without residue retention; RTR = Reduced tillage with residue retention, RTWR = Reduced tillage without residue retention, CTR = Conventional tillage with residue incorporation; CT = Conventional tillage without residue incorporation.

the conjoint application of INF + FYM (0.82%) and sole application of INF (0.65%) than control (0.54%). Kumar et al.^[83] reported that the CT without residue retention had significantly lower labile carbon fractions (27%–48%) than zero-tillage with 6-ton residue retention. Moreover, residue-retained fertilized treatments had significantly greater labile fractions of C than sole fertilized treatments^[125]. Kumar et al.^[83] reported highest change in DOC in zero-till with residue retention (28.2%) in comparison to conventional tillage practices. In ZT, absence of soil perturbations resulted in sustained supply of organic substrata for soil microbes which increases their activity. On the contrary, CT practices resulted in higher losses of C as CO₂ due to frequent disturbances.

Conclusions

The soil characteristics such as plant available nutrients, microbial diversity and soil organic carbon transformation are dwindling on account of intensive cultivation under conventional tillage practices, therefore, demand relevant management approaches for soil and crop sustainability. Long-term application of organic amendments significantly increases soil properties by increasing plant available macro, micro, secondary nutrients and soil organic C, whereas the increase in organic C by INF application is, by and large, due to increment in organic C content within macro-aggregates and in the silt + clay compartments. The soil organic carbon and other plant available nutrients were significantly greater in conservation tillage systems as compared to conventional tillage (CT) that conservation approaches could be an exemplary promoter of soil productivity by modifying soil structure thereby protecting SOM and maintaining higher nutrient content. The mean concentration of different fractions of carbon MBN, PMN and soil respiration under integrated nutrient management treatments was higher as compared with to control. Therefore, the conjoint use of organic manures or retention of crop residues with inorganic fertilizers is imperative to reduce the depletion of SOC while sustaining crop production as a realistic alternative. Future research should focus mainly on the usage of

organic and mineral fertilizers in conjunction with conservation tillage approaches to sustain the soil environment.

Author contributions

The authors confirm contribution to the paper as follows: study conception and design: Dhaliwal SS, Shukla AK, Randhawa MK, Behera SK; data collection: Sanddep S, Dhaliwal SS, Behera SK; analysis and interpretation of results: Dhaliwal SS, Gagandeep Kaur, Behera SK; draft manuscript preparation: Dhaliwal SS, walia, Shukla AK, Toor AS, Behera SK, Randhawa MK. All authors reviewed the results and approved the final version of the manuscript.

Data availability

Data sharing not applicable to this article as no datasets were generated or analyzed during the current study.

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Conflict of interest

The authors declare that they have no conflict of interest.

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References

1. Tahat MM, Alananbeh KM, Othman YA, Leskovar DI. 2020. Soil health and sustainable agriculture. *Sustainability* 12:4859

2. Celestina C, Hunt JR, Sale PWG, Franks AE. 2019. Attribution of crop yield responses to application of organic amendments: A critical review. *Soil and Tillage Research* 186:135–45
3. Tan ZX, Lal R, Wiebe KD. 2005. Global soil nutrient depletion and yield reduction. *Journal of Sustainable Agriculture* 26(1):123–46
4. Mikha MM, Obour AK, Holman JD. 2018. Soil nutrients status after fifty years of tillage and nitrogen fertilization. *Communications in Soil Science and Plant Analysis* 49:1953–75
5. Tiecher T, Calegari A, Caner L, Rheinheimer DdS. 2017. Soil fertility and nutrient budget after 23-years of different soil tillage systems and winter cover crops in a subtropical Oxisol. *Geoderma* 308:78–85
6. Ahmad AA, Radovich TJK, Nguyen HV, Uyeda J, Arakaki A, et al. 2016. Use of organic fertilizers to enhance soil fertility, plant growth, and yield in a tropical environment. In *Organic Fertilizers - From basic concepts to applied outcomes*, eds. Larramendy ML, Soloneski S. Rijeka: IntechOpen. <https://doi.org/10.5772/62529>
7. Srivastava P, Balhara M, Giri B. 2020. Soil Health in India: Past History and Future Perspective. In *Soil Health. Soil Biology*, eds. Giri B, Varma A. vol 59. Cham: Springer. pp. 1–19. https://doi.org/10.1007/978-3-030-44364-1_1
8. Schlöter M, Nannipieri P, Sørensen SJ, van Elsas JD. 2018. Microbial indicators for soil quality. *Biology and Fertility of Soils* 54:1–10
9. Paul EA. 2007. Soil microbiology, ecology and biochemistry in perspective. In *Soil Microbiology, Ecology and Biochemistry*, eds. Paul EA. 3rd Edition. San Diego: Academic Press. pp. 3–24. <https://doi.org/10.1016/b978-0-08-047514-1.50005-6>
10. Geisseler D, Scow KM. 2014. Long-term effects of mineral fertilizers on soil microorganisms – A review. *Soil Biology and Biochemistry* 75:54–63
11. Sollins P, Swanston C, Kramer M. 2007. Stabilization and destabilization of soil organic matter—a new focus. *Biogeochemistry* 85:1–7
12. Wu T, Schoenau JJ, Li F, Qian P, Malhi SS, et al. 2004. Influence of cultivation and fertilization on total organic carbon and carbon fractions in soils from the Loess Plateau of China. *Soil and Tillage Research* 77:59–68
13. Dhaliwal SS, Naresh RK, Mandal A, Singh R, Dhaliwal MK. 2019a. Dynamics and transformations of micronutrients in agricultural soils as influenced by organic matter build-up: A review. *Environmental and Sustainability Indicators* 1–2:100007
14. Ozlu E, Kumar S. 2018. Response of soil organic carbon, pH, electrical conductivity, and water stable aggregates to long-term annual manure and inorganic fertilizer. *Soil Science Society of America Journal* 82(5):1243–51
15. Šimon T. 2008. The influence of long-term organic and mineral fertilization on soil organic matter. *Soil and Water Research* 3:41–51
16. Weil RR, Islam KR, Stine MA, Gruver JB, Samson-Liebig SE. 2003. Active carbon for soil quality assessment: A simplified method for laboratory and field use. *American Journal of Alternative Agriculture* 18:3–17
17. Majumder B, Mandal B, Bandyopadhyay PK, Gangopadhyay A, Mani PK, et al. 2008. Organic amendments influence soil organic carbon pools and rice–wheat productivity. *Soil Science Society of America Journal* 72:775–85
18. Mandal A, Toor AS, Dhaliwal SS. 2020. Assessment of sequestered organic carbon and its pools under different agricultural land-uses in the semi-arid soils of south-western Punjab, India. *Journal of Soil Science and Plant Nutrition* 20:259–73
19. Wei K, Bao H, Huang S, Chen L. 2017. Effects of long-term fertilization on available P, P composition and phosphatase activities in soil from the Huang-Huai-Hai Plain of China. *Agriculture, Ecosystems & Environment* 237:134–42
20. Zhu L, Hu N, Yang M, Zhan X, Zhang Z. 2014. Effects of different tillage and straw return on soil organic carbon in a rice–wheat rotation system. *PLoS ONE* 9(2):e88900
21. Andruschkewitsch R, Koch HJ, Ludwig B. 2014. Effect of long-term tillage treatments on the temporal dynamics of water-stable aggregates and on macro-aggregate turnover at three German sites. *Geoderma* 217–218:57–64
22. Naresh RK, Bhaskar S, Dhaliwal SS, Kumar A, Gupta RK, et al. 2018a. Soil carbon and nitrogen mineralization dynamics following incorporation and surface application of rice and wheat residues in a semi-arid area of North West India: A review. *Journal of Pharmacognosy and Phytochemistry* 7(1):248–59
23. Naresh RK, Timsina J, Bhaskar S, Gupta RK, Singh AK, et al. 2017b. Effects of tillage, residue and nutrient management on soil organic carbon dynamics and its fractions, soil aggregate stability and soil carbon sequestration: A review. *EC Nutrition* 12(2):53–80
24. Liu ZP, Shao MA, Wang YQ. 2012. Estimating soil organic carbon across a large-scale region: A state-space modeling approach. *Soil Science* 177:607–18
25. Hartmann C, Poss R, Noble AD, Jongskul A, Bourdon E, et al. 2008. Subsoil improvement in a tropical coarse textured soil: effect of deep-ripping and slotting. *Soil and Tillage Research* 99(2):245–53
26. Dobermann A, Cassman KG. 2002. Plant nutrient management for enhanced productivity in intensive grain production systems of the United States and Asia. *Plant and Soil* 247:153–75
27. Yue X, Zhang J, Shi A, Yao S, Zhang B. 2016. Manure substitution of mineral fertilizers increased functional stability through changing structure and physiology of microbial communities. *European Journal of Soil Biology* 77:34–43
28. Han SH, An JY, Hwang J, Kim SB, Park BB. 2016. The effects of organic manure and chemical fertilizer on the growth and nutrient concentrations of yellow poplar (*Liriodendron Tulipifera* Lin.) in a nursery system. *Forest Science and Technology* 12(3):137–43
29. Hasnain M, Chen J, Ahmed N, Memon S, Wang L, et al. 2020. The effects of fertilizer type and application time on soil properties, plant traits, yield and quality of tomato. *Sustainability* 12(21):9065
30. Liu Z, Rong Q, Zhou W, Liang G. 2017. Effects of inorganic and organic amendment on soil chemical properties, enzyme activities, microbial community and soil quality in yellow clayey soil. *PLoS ONE* 12(3):e0172767
31. Shang L, Wan L, Zhou X, Li S, Li X. 2020. Effects of organic fertilizer on soil nutrient status, enzyme activity, and bacterial community diversity in *Leymus chinensis* steppe in Inner Mongolia, China. *PLoS ONE* 15(10):e0240559
32. Shiwakoti S, Zheljazkov VD, Gollany HT, Kleber M, Xing B, et al. 2020. Macronutrient in soils and wheat from long-term agroexperiments reflects variations in residue and fertilizer inputs. *Scientific Reports* 10:3263
33. Ge S, Zhu Z, Jiang Y. 2018. Long-term impact of fertilization on soil pH and fertility in an apple production system. *Journal of Soil Science and Plant Nutrition* 18(1):282–93
34. Ning CC, Gao PD, Wang BQ, Lin WP, Jiang NH, et al. 2017. Impacts of chemical fertilizer reduction and organic amendments supplementation on soil nutrient, enzyme activity and heavy metal content. *Journal of Integrative Agriculture* 16(8):1819–31
35. Chen D, Yuan L, Liu Y, Ji J, Hou H. 2017. Long-term application of manures plus chemical fertilizers sustained high rice yield and improved soil chemical and bacterial properties. *European Journal of Agronomy* 90:34–42
36. Shiwakoti S, Zheljazkov VD, Gollany HT, Kleber M, Xing B. 2019b. Macronutrients in soil and wheat as affected by a long-term tillage and nitrogen fertilization in winter wheat–fallow rotation. *Agronomy* 9:178
37. Dorneles EP, Lisboa BB, Abichequer AD, Bissani CA, Meurer EJ, et al. 2015. Tillage, fertilization systems and chemical attributes of a Paleudult. *Scientia Agricola* 72(2):175–86
38. Kumar J, De N, Pradhan A, Sharma P, Charry GR. 2018a. Effect of tillage on nutrient dynamics and physical condition of soil under

- rainfed rice chick pea-based cropping system of Bastar, Chhattisgarh. *International Journal of Current Microbiology and Applied Sciences* 7(12):2843–50
39. Shiwakoti S, Zheljazkov VD, Gollany HT, Kleber M, Xing B. 2019c. Effect of tillage on macronutrients in soil and wheat of a long-term dryland wheat-pea rotation. *Soil and Tillage Research* 190:194–201
 40. Li BY, Zhou DM, Cang L, Zhang HL, Fan XH, et al. 2007. Soil micronutrient availability to crops as affected by long-term inorganic and organic fertilizer applications. *Soil and Tillage Research* 96(1–2):166–73
 41. Li BY, Huang SM, Wei MB, Zhang HL, Shen AL, et al. 2010. Dynamics of soil and grain micronutrients as affected by long-term fertilization in an Aquic Inceptisol. *Pedosphere* 20(6):725–35
 42. López-Fando C, Pardo MT. 2009. Changes in soil chemical characteristics with different tillage practices in a semi-arid environment. *Soil and Tillage Research* 104(2):278–84
 43. Martín-Rueda I, Muñoz-Guerra LM, Yunta F, Esteban E, Tenorio JL, et al. 2007. Tillage and crop rotation effects on barley yield and soil nutrients on a *Calcitridic Haploxeralf*. *Soil and Tillage Research* 92(1–2):1–9
 44. Sharma S, Dhaliwal SS. 2020. Rice residue incorporation and nitrogen application: effects on yield and micronutrient transformations under rice–wheat cropping system. *Journal of Plant Nutrition* 43(8):2697–711
 45. Sharma S, Dhaliwal SS. 2019. Effect of sewage sludge and rice straw compost on yield, micronutrient availability and soil quality under rice–wheat system. *Communications in Soil Science and Plant Analysis* 50(16):1943–54
 46. Dhaliwal SS, Dhaliwal MK, Shukla AK, Manchanda JS. 2017. Long-term effect of integrated nutrient management on distribution of DTPA-extractable micronutrients in Typic Haplustep under rice–wheat system. *Indian Journal of Fertilizers* 13(8):54–57
 47. Dhaliwal SS, Naresh RK, Walia MK, Gupta RK, Mandal A, et al. 2020. Long-term effects of intensive rice–wheat and agroforestry based cropping systems on build-up of nutrients and budgets in alluvial soils of Punjab, India. *Archives of Agronomy and Soil Science* 66:330–42
 48. Dhaliwal SS, Walia MK, Phutela RP. 2012. Effect of inorganic fertilizers and manures application on macro and micronutrients distribution under long term rice–wheat system. *Journal of Plant Science Research* 28(1):149–61
 49. Shiwakoti S, Zheljazkov VD, Gollany HT, Kleber M, Xing B. 2019a. Micronutrients decline under long-term tillage and nitrogen fertilization. *Scientific Reports* 9:12020
 50. Lozano-García B, Parras-Alcántara L. 2014. Changes in soil properties and soil solution nutrients due to conservation versus conventional tillage in Vertisols. *Archives of Agronomy and Soil Science* 60(10):1429–44
 51. Alam MK, Islam MM, Salahin N, Hasanuzzaman M. 2014. Effect of tillage practices on soil properties and crop productivity in wheat-mungbean-rice cropping system under subtropical climatic conditions. *The Scientific World Journal* 2014:437283
 52. Dhaliwal SS, Singh BS. 2013. Depthwise distribution of macronutrients, micronutrients and microbial population in under different land use systems. *An Asian Journal of Soil Science* 8(2):404–11
 53. Johnson NC, Stefan S, Kelly S R, Philippe L, Paul E, et al. 2016. *Global Soil Biodiversity Atlas*. Luxembourg: European Commission, Publications Office of the European Union. 180 pp. <https://doi.org/10.2788/2613>
 54. Thiele-Bruhn S, Bloem J, de Vries FT, Kalbitz K, Wagg C. 2012. Linking soil biodiversity and agricultural soil management. *Current Opinion in Environmental Sustainability* 4:523–28
 55. Dangi S, Gao S, Duan Y, Wang D. 2020. Soil microbial community structure affected by biochar and fertilizer sources. *Applied Soil Ecology* 150:103452
 56. Luan H, Gao W, Huang S, Tang J, Li M, et al. 2020. Substitution of manure for chemical fertilizer affects soil microbial community diversity, structure and function in greenhouse vegetable production systems. *PLoS ONE* 15:e0214041
 57. Lazcano C, Gómez-Brandón M, Revilla P, Domínguez J. 2013. Short-term effects of organic and inorganic fertilizers on soil microbial community structure and function: A field study with sweet corn. *Biology and Fertility of Soils* 49:723–33
 58. Kamaa M, Mburu H, Blanchart E, Chibole L, Chotte JL, et al. 2011. Effects of organic and inorganic fertilization on soil bacterial and fungal microbial diversity in the Kabete long-term trial, Kenya. *Biology and Fertility of Soils* 47:315–21
 59. Liu L, Li C, Zhu S, Xu Y, Li H, et al. 2020. Combined application of organic and inorganic nitrogen fertilizers affects soil prokaryotic communities compositions. *Agronomy* 10(1):132
 60. Li Yu-mei, Duan Y, Wang G, Wang A, Shao G, et al. 2021. Straw alters the soil organic carbon composition and microbial community under different tillage practices in a meadow soil in North-east China. *Soil and Tillage Research* 208:104879
 61. He L, Lu S, Wang C, Mu J, Zhang Y, et al. 2021. Changes in soil organic carbon fractions and enzyme activities in response to tillage practices in the Loess Plateau of China. *Soil and Tillage Research* 209:104940
 62. Jenkinson DS, Ladd JN. 1981. Microbial biomass in soil: Measurement and turnover. In *Soil Biochemistry*, eds. Paul EA, Ladd JN. New York, USA: Marcel Dekker. pp. 415–17. www.scirp.org/reference/referencespapers?referenceid=1664759
 63. Dhaliwal SS. 2008. Profile distribution of chemical, physical and biological indicators in different land use systems under Takarala watershed in submontaneous tract of Punjab. *Journal of Plant Science and Research* 24(2):141–50
 64. Zhang JS, Guo JF, Chen GS, Qian W. 2005. Soil microbial biomass and its controls. *Journal of Forestry Research* 16:327–30
 65. Fließbach A, Oberholzer HR, Gunst L, Mäder P. 2007. Soil organic matter and biological soil quality indicators after 21 years of organic and conventional farming. *Agriculture, Ecosystems & Environment* 118:273–84
 66. Dou F, Wright AL, Hons FM. 2008. Sensitivity of labile soil organic carbon to tillage in wheat-based cropping systems. *Soil Science Society of America Journal* 72:144–53
 67. Liang B, Yang X, He X, Zhou J. 2011. Effects of 17-year fertilization on soil microbial biomass C and N and soluble organic C and N in loessial soil during maize growth. *Biology and Fertility of Soils* 47(2):121–28
 68. Mandal N, Dwivedi BS, Meena MC, Singh D, Datta SP, et al. 2013. Effect of induced defoliation in pigeonpea, farmyard manure and sulphitation pressmud on soil organic carbon fractions, mineral nitrogen and crop yields in a pigeonpea–wheat cropping system. *Field Crops Research* 154:178–87
 69. Dhaliwal SS, Naresh RK, Mandal A, Walia MK, Gupta RK, et al. 2019b. Effect of manures and fertilizers on soil physical properties, build-up of macro and micronutrients and uptake in soil under different cropping systems: A review. *Journal of Plant Nutrition* 42(20):2873–900
 70. Malviya SR. 2014. *Effect of conservation agricultural practices on selected soil physical properties and carbon pools in black soils of central India*. M. Sc. Thesis. Jawaharlal Nehru Krishi Vishwa Vidyalyaya, Jabalpur, Madhya Pradesh.
 71. Naresh RK, Gupta RK, Vivek, Rathore RS, Singh SP, et al. 2018b. Carbon, nitrogen dynamics and soil organic carbon retention potential after 18 years by different land uses and nitrogen management in RWCS under Typic Ustochrept soil. *International Journal of Current Microbiology and Applied Sciences* 7(12):3376–99
 72. Nakhro N, Dkhar MS. 2010. Impact of organic and inorganic fertilizers on microbial populations and biomass carbon in paddy field soil. *Journal of Agronomy* 9(3):102–10

73. Ramdas MG, Manjunath BL, Singh NP, Ramesh R, Verma RR, et al. 2016. Effect of organic and inorganic sources of nutrients on soil microbial activity and soil organic carbon build-up under rice in west coast of India. *Archives of Agronomy and Soil Science* 63:414–26
74. Xu Y, Tang H, Xiao X, Li W, Li Chao. et al. 2018. Effects of long-term fertilization management practices on soil microbial carbon and microbial biomass in paddy soil at various stages of rice growth. *Revista Brasileira de Ciência do Solo* 42:e0170111
75. Xiao SS, Ye YY, Xiao D, Chen WR, Zhang W, et al. 2019. Effects of tillage on soil N availability, aggregate size, and microbial biomass in a subtropical karst region. *Soil and Tillage Research* 192:187–95
76. Li Y, Li Z, Cui S, Liang G, Zhang Q. 2021. Microbial-derived carbon components are critical for enhancing soil organic carbon in no-tillage croplands: A global perspective. *Soil and Tillage Research* 205:104758
77. Dhaliwal SS, Sharma BD, Singh BS, Khera KL. 2009. Profile distribution of microbial parameters, macro and micronutrients in four land use systems of Kular watershed in submontaneous tract of Punjab. *Environment and Ecology* 27(1):11–17
78. Černý J, Balík J, Kulhánek M, Nedvěď V. 2008. The changes in microbial biomass C and N in long-term field experiments. *Plant, Soil and Environment* 54(5):212–18
79. El-Sharkawi HM. 2012. Effect of nitrogen sources on microbial biomass nitrogen under different soil types. *International Scholarly Research Notices* 2012:310727
80. Sugihara S, Funakawa S, Kilasara M, Kosaki T. 2010. Dynamics of microbial biomass nitrogen in relation to plant nitrogen uptake during the crop growth period in a dry tropical cropland in Tanzania. *Soil Science and Plant Nutrition* 56:105–14
81. Malik MA, Khan KS, Marschner P, Fayyaz-ul-Hassan. 2013. Microbial biomass, nutrient availability and nutrient uptake by wheat in two soils with organic amendments. *Journal of Soil Science and Plant Nutrition* 13(4):955–66
82. Wang Y, Zhang MH, Zhao XL. 2019. Effects of organic amendments on microbial biomass carbon and nitrogen uptake by corn seedlings grown in two purple soils. *Environmental Science* 40(8):3808–15
83. Kumar V, Naresh RK, Kumar S, Kumar S, Kumar S, et al. 2018b. Tillage, crop residue, and nitrogen levels on dynamics of soil labile organic carbon fractions, productivity and grain quality of wheat crop in Typic Ustochrept soil. *Journal of Pharmacognosy and Phytochemistry* 7(1):598–609
84. Ginakes P, Grossman JM, Baker JM, Dobbratz M, Sooksan-guan T. 2018. Soil carbon and nitrogen dynamics under zone tillage of varying intensities in a kura clover living mulch system. *Soil and Tillage Research* 184:310–16
85. Aulakh MS, Garg AK, Kumar S. 2013. Impact of integrated nutrient, crop residue and tillage management on soil aggregates and organic matter fractions in semiarid subtropical soil under soybean-wheat rotation. *American Journal of Plant Sciences* 4:2148–64
86. Verma N, Goyal S. 2018. Effect of organic manuring and integrated nutrient management on microbiological properties in sandy soils under wheat-guar cropping system. *International Journal of Current Microbiology and Applied Sciences* 7:1340–51
87. Mahal NK, Castellano MJ, Miguez FE. 2018. Conservation agriculture practices increase potentially mineralizable nitrogen: A meta-analysis. *Soil Science Society of America Journal* 82:1270–78
88. Baker JM, Ochsner TE, Venterea RT, Griffis TJ. 2007. Tillage and soil carbon sequestration- What do we really know? *Agriculture Ecosystems and Environment* 118:1–5
89. Raich JW, Schlesinger WH. 1992. The global carbon dioxide flux in soil respiration and its relationship to vegetation and climate. *Tellus B: Chemical and Physical Meteorology* 44(2):81–99
90. Gilani SS, Bahmanyar MA. 2008. Impact of organic amendments with and without mineral fertilizers on soil microbial respiration. *Journal of Applied Sciences* 8:642–47
91. Celis JE, Sandoval M, Martínez B, Quezada C. 2013. Effect of organic and mineral amendments upon soil respiration and microbial biomass in a saline-sodic soil. *Ciencia e Investigación Agraria* 40:571–80
92. Faust S, Koch HJ, Joergensen RG. 2019. Respiration response to different tillage intensities in transplanted soil columns. *Geoderma* 352:289–97
93. Bongiorno G, Bünemann EK, Brussaard L, Mäder P, Oguejiofor CU, et al. 2020. Soil management intensity shifts microbial catabolic profiles across a range of European long-term field experiments. *Applied Soil Ecology* 154:103596
94. Kalkhaje YK, He Z, Yang X, Lu Y, Zhou J, et al. 2021. Co-application of nitrogen and straw-decomposing microbial inoculant enhanced wheat straw decomposition and rice yield in a paddy soil. *Journal of Agriculture and Food Research* 4:100134
95. Gong H, Li J, Ma J, Li F, Ouyang Z, et al. 2018. Effects of tillage practices and microbial agent applications on dry matter accumulation, yield and the soil microbial index of winter wheat in North China. *Soil and Tillage Research* 184:235–42
96. FAO. 2015. *Revised world soil charter*. Rome: Food and Agriculture Organization of the United Nations. www.fao.org/documents/card/en/c/e60df30b-0269-4247-a15f-db564161fee0/
97. Kane D. 2015. Carbon sequestration potential on agricultural lands: A review of current science and available practices. *National Sustainable Agriculture Coalition Breakthrough Strategies and Solutions, LLC*. pp. 1–35. https://sustainableagriculture.net/wp-content/uploads/2015/12/Soil_C_review_Kane_Dec_4-final-v4.pdf
98. Katkar RN, Kharche VK, Sonune BA, Wanjari RH, Singh M. 2012. Long-term effect of nutrient management on soil quality and sustainable productivity under sorghum-wheat crop sequence in Vertisol of Akola. *Agropedology* 22(2):103–14
99. Mazumdar SP, Kundu DK, Nayak AK, Ghosh D. 2015. Soil aggregation and associated organic carbon as affected by long term application of fertilizer and organic manures under rice-wheat system in middle Gangetic plains of India. *Journal of Agricultural Physics* 15(2):113–21
100. Dhaliwal SS, Ram H, Walia SS, Walia MK, Kumar B, et al. 2019d. Long-term influence of nutrient management on carbon and nutrients in Typic-Ustochrept soils. *Communications in Soil Science and Plant Analysis* 50:2420–28
101. Datta A, Jat HS, Yadav AK, Choudhary M, Sharma PC, et al. 2019. Carbon mineralization in soil as influenced by crop residue type and placement in an Alfisols of Northwest India. *Carbon Management* 10:37–50
102. Naresh RK, Kumar A, Bhaskar S, Dhaliwal SS, Vivek, et al. 2017a. Organic matter fractions and soil carbon sequestration after 15-years of integrated nutrient management and tillage systems in an annual double cropping system in northern India. *Journal of Pharmacognosy and Phytochemistry* 6(6):670–83
103. Cai A, Xu H, Shao X, Zhu P, Zhang W, et al. 2016. Carbon and nitrogen mineralization in relation to soil particle-size fractions after 32 years of chemical and manure application in a continuous maize cropping system. *PLoS ONE* 11(3):e0152521
104. Zhao H, Shar AG, Li S, Chen Y, Shi J, et al. 2018. Effect of straw return mode on soil aggregation and aggregate carbon content in an annual maize-wheat double cropping system. *Soil and Tillage Research* 175:178–86
105. Srinivasarao C, Venkateswarlu B, Lal R, Singh AK, Kundu S, et al. 2014. Long-term manuring and fertilizer effects on depletion of soil organic carbon stocks under pearl millet-cluster bean-castor rotation in Western India. *Land Degradation & Development* 25(2):173–83

106. Srinivasarao C, Venkateswarlu B, Lal R, Singh AK, Vittal KPR, et al. 2012. Long-term effects of soil fertility management on carbon sequestration in a rice–lentil cropping system of the Indo-Gangetic Plains. *Soil Science Society of America Journal* 76(1):168–78
107. Tong X, Xu M, Wang X, Bhattacharyya R, Zhang W, et al. 2014. Long-term fertilization effects on organic carbon fractions in a red soil of China. *Catena* 113:251–59
108. Naresh RK, Gupta RK, Jat ML, Singh SP, Dwivedi A, et al. 2016. Tillage, irrigation levels and rice straw mulches effects on wheat productivity, soil aggregates and soil organic carbon dynamics after rice in sandy loam soils of subtropical climatic conditions. *Journal of Pure and Applied Microbiology* 10(3):1061–2002
109. Zhang X, Sun N, Wu L, Xu M, Bingham IJ, et al. 2016. Effects of enhancing soil organic carbon sequestration in the topsoil by fertilization on crop productivity and stability: Evidence from long-term experiments with wheat–maize cropping systems in China. *Science of the Total Environment* 562:247–59
110. Choudhury SG, Yaduvanshi NPS, Chaudhari SK, Sharma DR, Sharma DK, et al. 2018. Effect of nutrient management on soil organic carbon sequestration, fertility, and productivity under rice–wheat cropping system in semi-reclaimed sodic soils of North India. *Environment Monitoring and Assessment* 190:117
111. Franzluebbers AJ, Hubbs MD, Norfleet ML. 2012. Evaluating soil organic carbon sequestration potential in the Cotton Belt with the soil conditioning index. *Journal of Soil and Water Conservation* 67:378–89
112. Mandal KG, Baral U, Padhi J, Majhi P, Chakraborty H, et al. 2012. Effects of cropping on soil properties and organic carbon stock in Deras region, India. *Regional Environmental Change* 12(4):899–912
113. Jyoti Nath A, Bhattacharyya T, Deka J, Kumar Das A, Ray SK. 2015. Management effect on soil organic carbon pools in lowland rain-fed paddy growing soil. *Journal of Tropical Agriculture* 53(2):131–38
114. Das D, Dwivedi BS, Singh VK, Datta SP, Meena MC, et al. 2017. Long-term effects of fertilisers and organic sources on soil organic carbon fractions under a rice–wheat system in the Indo-Gangetic Plains of north-west India. *Soil Research* 55(3):296–308
115. Fang XM, Chen FS, Wan SZ, Yang QP, Shi JM. 2015. Topsoil and deep soil organic carbon concentration and stability vary with aggregate size and vegetation type in subtropical China. *PLoS ONE* 10(9):e0139380
116. Pandey D, Agrawal M, Bohra JS, Adhya TK, Bhattacharyya P. 2014. Recalcitrant and labile carbon pools in a sub-humid tropical soil under different tillage combinations: A case study of rice–wheat system. *Soil and Tillage Research* 143:116–22
117. Xue JF, Pu C, Liu SL, Chen ZD, Chen F, et al. 2015. Effects of tillage systems on soil organic carbon and total nitrogen in a double paddy cropping system in Southern China. *Soil and Tillage Research* 153:161–68
118. Singh A, Phogat VK, Dahiya R, Batra SD. 2014. Impact of long-term zero till wheat on soil physical properties and wheat productivity under rice–wheat cropping system. *Soil and Tillage Research* 140:98–105
119. Kuhn NJ, Hu Y, Bloemertz L, He J, Li H, et al. 2016. Conservation tillage and sustainable intensification of agriculture: regional vs. global benefit analysis. *Agriculture, Ecosystems & Environment* 216:155–65
120. Sapkota TB, Jat RK, Singh RG, Jat ML, Stirling CM, et al. 2017. Soil organic carbon changes after seven years of conservation agriculture in a rice–wheat system of the eastern Indo-Gangetic Plains. *Soil Use and Management* 33:81–89
121. Mandal A, Toor AS, Dhaliwal SS. 2018. Effect of land-uses on physico-chemical properties and nutrient status of surface (0–15 cm) and sub-surface (15–30 cm) layers in soils of south-western Punjab, India. *International Journal of Current Microbiology and Applied Sciences* 7(6):2659–71
122. Banger K, Toor GS, Biswas A, Sidhu SS, Sudhir K. 2010. Soil organic carbon fractions after 16-years of applications of fertilizers and organic manure in a *Typic Rhodalfs* in semi-arid tropics. *Nutrient Cycling in Agroecosystems* 86:391–99
123. Nayak AK, Gangwar B, Shukla AK, Mazumdar SP, Kumar A, et al. 2012. Long-term effect of different integrated nutrient management on soil organic carbon and its fractions and sustainability of rice–wheat system in Indo Gangetic Plains of India. *Field Crops Research* 127:129–39
124. Gu C, Liu Y, Mohamed I, Zhang R, Wang X, et al. 2016. Dynamic changes of soil surface organic carbon under different mulching practices in citrus orchards on sloping land. *PLoS ONE* 11(12):e0168384
125. Naresh RK, Gupta RK, Pal G, Dhaliwal SS, Kumar D, et al. 2015. Tillage crop establishment strategies and soil fertility management: Resource use efficiencies and soil carbon sequestration in a rice–wheat cropping system. *Ecology, Environment and Conservation* 21:S121–S128
126. Anantha KC, Majumder SP, Padhan D, Badole S, Datta A, et al. 2018. Carbon dynamics, potential and cost of carbon sequestration in double rice cropping system in semi-arid southern India. *Journal of Soil Science and Plant Nutrition* 18(2):418–34
127. Wang W, Lai DYF, Wang C, Pan T, Zeng C. 2015. Effects of rice straw incorporation on active soil organic carbon pools in a subtropical paddy field. *Soil and Tillage Research* 152:8–16
128. Tiwari R, Naresh RK, Vivek, Jat L, Purushattom, et al. 2018. Soil aggregation and aggregate associated organic carbon fractions and microbial activities as affected by tillage and straw management in a rice–wheat rotation: A review. *Journal of Pharmacognosy and Phytochemistry* 7(5):2865–93



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